Calibration and performances of in-situ gamma ray spectrometer

by

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Abstract

Since 1896, when Henri Becquerel discovered that penetrating radiation was given off in the radioactive decay of uranium, the studies on radioactivity have been an interest of scientific world. With the spread of nuclear technologies applied to energy, health and industrial production, the theme of environmental radioactivity monitoring increasingly is becoming important to the policies of the health public protection both national and European level. Italy is required to comply with the recommendation of the European Commission of 8 June 2000 on the application of Article 36 of the Euratom Treaty concerning the monitoring of levels of radioactivity in the environment for the purpose of assessing the exposure of the population as a whole. In addition, the World Health Organization has identified the first group of carcinogens gas $^{222}$Rn, which is considered the second leading cause, after smoking, of lung tumors. In our environment there are various sources of radioactivity that can be natural or artificial origin. Gamma-ray spectrometry is a widely used and powerful method that can be employed both to identify and quantify radionuclides. The purpose of this work is calibration and performances of in situ a portable gamma ray spectrometer.

In the first chapter I have given the necessary concepts for understanding the phenomenon of radioactivity. Qualitatively has been described the process of radioactive decay and its three types which can occur in nature. Three categories of environmental radionuclides, cosmogenic, primordial and man-made are discussed. We are exposed to environmental radiation from different sources. The origin of radioactivity in the environment can be divided into two main sources: (a) natural and (b) man-made sources. Mostly the naturally occurring radiation arises from terrestrial radioactive nuclides that are widely distributed in the earth’s crust and extra-terrestrial sources arising from cosmic ray. Also from human activities arise some other sources concerned with the use of radiation and radioactive materials from which releases of radionuclides into the environment may occur.

In the second chapter is described the gamma radiation interacts with matter via three main processes: the photoelectric effect, Compton scattering and pair production. The operation of a detector is based on the interaction of photons constituting the incident radiation with the material that constitutes the detector itself. Thanks to these processes, all or part of the energy possessed by the radiation is transferred to the mass of the detector
and then converted into an electrical signal. The basic notions related to the interaction of electromagnetic radiation with matter that we will provide in this chapter will therefore be useful to understand the mechanisms that are at the basis of the generation of a gamma spectrum. In addition, this chapter will briefly describe the two main types of gamma radiation detectors, i.e. the semiconductor detector and the scintillation, in particular the high-pure germanium detector (HPGe) and a sodium iodide detector activated by thallium NaI(Tl).

In the third chapter is described the study area in which are performed the measurements of natural radioactivity. The area under consideration is the Ombrone basin located in southern Tuscany and Commune of Schio located in Region of Veneto. During the campaign were acquired in situ 338 spectra, including 80 with the ZaNaI_1.0L placed on the ground (Ombrone), 80 with the ZaNaI_1.0L placed on a tripod at 1m height (Ombrone), 89 with the ZaNaI_1.0L placed on the ground (Schio) and 89 spectra are acquired with a backpack placed on the shoulders of an operator (Schio). In each of the 80 sites which have been realized the measurements of radioactivity with the ZaNaI_1.0L instrument, also have been taken 5 different soil samples, for a total of 400 samples. The abundances of 40K, 238U and 232Th were obtained from the analysis of 338 spectra taken with the ZaNaI_1.0L and 400 spectra measured on soil samples in the laboratory with a high-pure germanium detector (MCA_Rad). Also it is described the procedures of ZaNaI_1.0L portable scintillation gamma-ray spectrometers for in-situ measurements.

In the fourth chapter is described the procedure for the preparation of soil samples to be analyzed with the MCA_Rad system. The gamma-ray spectrometry system, called MCA_Rad introduces an innovative configuration of a laboratory high-resolution gamma-ray spectrometer featured with a complete automation measurement process, which can conduct measurements on each type of material (solid, liquid or gaseous) in less than 1 hour. The utilization of two coupled HPGe detectors permits to achieve good statistical accuracies in shorter time, which contributes in drastically reducing costs and man power involved. It is made a description of the characterization of absolute full-energy peak efficiency of such instrument reported here.

In the fifth chapter are discussed the correlations between the abundances of 40K, 238U and 232-Th measured with the ZaNaI_1.0L and those obtained from laboratory analysis on soil samples. The analysis was focused in particular on the study of four different types of correlation: correlation between in-situ acquisition on ground and laboratory
measurements, correlation between in-situ acquisition on tripod and laboratory measurements, correlation between in-situ acquisition on ground and on tripod and correlation between in-situ acquisition on ground and on operator shoulder and the influence of vegetative cover during measurements in-situ.
Keywords

Gamma-ray spectrometry; Environmental radioactivity monitoring; Gamma-ray spectrometry efficiency calibration; Semiconductor HPGe detector; Scintillation NaI(Tl) detector; In-situ gamma-ray spectrometry; Full spectrum analysis;
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Chapter 1

1. Introduction to environmental radioactivity

In this chapter are introduced the necessary concepts for understanding the phenomenon of radioactivity. Qualitatively has been described the process of radioactive decay and its three types which can occur in nature. The statistical nature of radioactive decay and the various natural radiation sources with respect to an artificial radioactivity are described.

1.1 Law of radioactive decay

In general, each nuclear reaction is associated with an amount of energy that takes the name of Q value, whereby it is possible to define if the reaction is exothermic or endothermic type. The Q value is defined by the difference between sum of initial masses and the sum of final masses:

\[ Q = \sum_i m_i - \sum_f m_f \]  

(Eq. 1.1)

If the value of Q is positive, the reaction is exothermic. An exothermic reaction occurs spontaneously and the final particles produced in the reaction divide a quantity of kinetic energy, which in the case when initial particles are at rest equals the Q value of the same reaction. Instead, the reaction is endothermic in the case when the Q value is negative, or in other words when the sum of final masses is greater than the sum of initial masses. Then the reaction can occur only in the case in which the initial particles are in motion and possess an amount of energy higher than the threshold value that makes possible the reaction. The radioactive decay is always an exothermic process, in which the energy is released in the form of radiation and by emitting particles. The radioactive decay is based on the fact that the decay, i.e. the transition of a parent nucleus to a daughter nucleus is a statistical process. The disintegration (decay) probability is a fundamental property of an atomic nucleus and remains equal in time. The law of radioactive decay of a given radioactive substance predicts how the number of nuclei \( N_0 \),
which are present at time $t_0$ decreases with time $t$. The number $dN$, decaying in a time interval $dt$, is proportional to $N$, therefore:

$$\frac{dN(t)}{dt} = -\lambda N(t) \quad \text{(Eq. 1.2)}$$

where $\lambda$ is the decay constant which equals the probability per unit time of the decay of an atom. The negative sign indicates that the number of radioactive nuclei decreases when the time increases. From the solution of the differential equation (Eq. 1.2) the exponential law of radioactive decay can be expressed as:

$$N(t) = N_0 e^{-\lambda t} \quad \text{(Eq. 1.3)}$$

where $N(t)$ is the number nuclei at present time $t$ and $N_0$ is the original number of nuclei at time $t_0 = 0$.

The half-life time of radionuclide $t_{1/2}$, in which the original number of the atoms it is reduced to one-half, it is used often to describe a radioactive decay. The half-life differs for different radionuclides and varies between few seconds to billions of years and expressed by equation:

$$t_{1/2} = \frac{\ln(2)}{\lambda} = \frac{0.693}{\lambda} \quad \text{(Eq. 1.4)}$$

where the parameter $1/\lambda$ is defined as the mean lifetime $\tau$, which is the average time that a nucleus is likely to survive before it decays. Since the activity, $A$, is proportional with the number of atoms present, it follows the same rate of decrease and can be obtained by differentiating the Eq. 1.3; i.e.,

$$A(t) = \left[ \frac{dN(t)}{dt} \right] = -\lambda N(t) = -\lambda N_0 e^{-\lambda t} \quad \text{(Eq. 1.5)}$$

$$A(t) = A_0 e^{-\lambda t} \quad \text{(Eq. 1.6)}$$
where $A_0$ is the initial activity at $t_0 = 0$. The SI unit of the activity defined as one disintegration per second is called Becquerel (Bq). Another unit of activity is the Curie (Ci) which is defined based on the activity of 1 gram of Radium ($^{226}$Ra) and is equal to 1 Ci = $3.7 \times 10^{10}$ Bq [Turner J.E., 2007]. Figure 1.1 shows how the activity changes with time following the exponential law of radioactivity.

![Figure 1.1: exponential decay of activity.](image)

Therefore, the time trend of a number of parent atoms decreases following the exponential law, while the number of daughter atoms increases follows an exponential trend as shown in figure 1.2.

![Figure 1.2: temporal trend of the number of parent atoms and daughter atoms in a radioactive sample.](image)

Some radionuclides may have more than one branches of decay, but independent from the mode of decay, their half-life is the same. For each branch of decay is associated a well-defined probability, represented by the quantity that takes the name of branching ratio. It corresponds to the fraction of particles that decay according to a certain decay
mode, compared to the total number of particles of the radioactive sample. The branching ratio is a very important property in the study of decays, because for each transition is released a finite amount of energy which is characteristic of the specific branch.

Through the detection of energies associated with the various processes of transition is possible to construct the overall energy spectrum, which allow to identify the type of decay occurred and the elements that produced them, could then go up to the likely progenitor nucleus. The radioactive decay often occurs in a series, or decay chain, in which the generated nuclides are radioactive; the chain ends at the moment when it reaches a stable nuclide. Each of the possible decays is characterized by the decay constant $\lambda_1, \lambda_2, \ldots, \lambda_i$ that expresses the probability that the specific process takes place in the unit of time. The result is that the overall system will be subject to decay with a total probability connected to constant $\lambda_t$ equal to the sum of the different decay constants, i.e.

$$\lambda_t = \lambda_1 + \lambda_2 + \cdots + \lambda_n \quad \text{(Eq. 1.7)}$$

In a closed system, starting from a given quantity of parent atoms, the number of daughter nuclides grows gradually until the moment in which the equilibrium is established within the decay chain. The condition of equilibrium is reached at the time in which the activities of all radionuclides of the decay chain are the equal:

$$\lambda_1 N_1 = \lambda_2 N_2 = \cdots = \lambda_i N_i \quad \text{(Eq. 1.8)}$$

This means that the number of daughter nuclides created in the unit of time corresponds exactly to the number of parent nuclides that decay per unit of time: in the case of a series this denotes that the unstable transition will decay with the same rate with which they are produced.
Figure 1.3: qualitative representation of the secular equilibrium concept. The atoms of $^{238}$U subject to decay are represented as a fluid which is poured into the container of $^{234}$Th. In secular equilibrium conditions the outgoing flow from the container of $^{234}$Th, corresponding to the number of atoms of $^{234}$Th subject to decay, will be equal to the incoming flow, corresponding to the number of atoms of $^{234}$Th produced by the decay of $^{238}$U. If the decay chain of $^{238}$U is in secular equilibrium in its entirety, equality between the incoming and the outcoming flow is valid for every element in the chain.

In this way the measurement of the concentration of one of the daughter nuclide can be used to estimate the concentration of any of the other elements belonging to the decay series, and in particular that of the parent nucleus. Inside a decay chain is possible to have a state of disequilibrium: this occurs when the parent nuclide of the chain has a half-life less than those of its progenies, and then disappears while the rest of the chain is active, or in the case in which the chain is broken, that is, when one of the daughter is separated and isolated from the chain.

1.2 The statistical nature of radioactive decay

The process of radioactive decay is a statistical phenomenon. Each disintegration that occurs during a radioactive decay is completely independent from the others, and the time interval between one decay and the next is not constant. For a high number of atoms of a given radionuclide which decay randomly, the frequency of radioactive decay is given by the Poisson distribution: if we denote with $\bar{n}$ the average rate of decay, with $P$ the probability that a certain number of nuclei $n$ decay at unit of time is given by:

$$P(n) = \frac{\bar{n}^n}{n!} e^{-\bar{n}} \hspace{1cm} (Eq. 1.9)$$
For a Poisson distribution, the variance $\sigma^2$ is equal to the mean value of the distribution, where $\sigma$ is the standard deviation. By increasing the average value of Poisson distribution tends to the Gaussian distribution, ideally for an infinite mean value the two distributions coincide. In this limit the distance of $1\sigma$ from the mean value expresses the confidence to find 68.3% of measures with respect to the total, the 95.5% within $2\sigma$ and 99.7% within $3\sigma$. In figure 1.4 are represented three Poisson distributions having different average value.

![Figure 1.4](image)

Figure 1.4: example of three Poisson distributions with mean value equal to 1, 5 and 10. It is observed that as average value increase, the Poisson distribution is close to Gauss ones.

If $N$ is the number of decay events recorded at the time $t$, then the standard deviation in the units of counts is given by:

$$\sigma(N) = \sqrt{N} \quad \text{(Eq. 1.10)}$$

where $\overline{N}$ is the expectation value of the number of counts (i.e., the average value of count assessed by carrying out a certain number of measures).

The relative standard deviation is expressed as:

$$\frac{\sigma(N)}{\overline{N}} = \frac{1}{\sqrt{\overline{N}}} \quad \text{(Eq. 1.11)}$$

By applying the propagation law of errors, for a rate of counts $n = \frac{N}{t}$ it is obtained the standard deviation i.e.,
Therefore the relative standard deviation is:

$$\frac{\sigma(n)}{n} = \frac{1}{\sqrt{nt}}$$  \hspace{1cm} \text{(Eq. 1.13)}$$

From equations above it can be observed that the accuracy of a radiometric measurement can be increased by increasing the number of counts N. This can be achieved by using an instrumentation more sensitive and efficient, improving the geometry of the measurement, or for a longer acquisition time.

1.3 Radioactive decay and decay modes

In nature there are stable nuclei until the lead, around which exists a zone of unstable nuclei, or radioactive, which spontaneously decay into neighbors nuclei with the emission of particles and energy. The stability of a nucleus is determined by its binding energy $E_b$, which corresponds to the energy required to split the nucleus into individual protons and neutrons, bringing them to a distance for which there is no strong nuclear interaction between them. This energy depends on the number of protons in relation to the number of neutrons: if the number of neutrons differ a lot from the number of protons the nucleus is unstable. We can define the binding energy per nucleon, $\varepsilon$, as ratio between the binding energy of nucleus and its mass number:

$$\varepsilon = \frac{E_b}{A}$$  \hspace{1cm} \text{(Eq. 1.14)}$$

The binding energy per nucleon represents the amount of energy required to split a nucleon (proton or neutron) at a nucleus of mass A.
Figure 1.5: binding energy per nucleon of common isotopes.

From figure 1.5 it can be seen that ε increase with A, up to A < 60: in correspondence of the iron reaches a maximum equal to 8.79 MeV, then decreases slowly. This decrease in binding energy beyond iron is due to the fact that, as a nucleus gets bigger, the ability of the strong force to counteract the electrostatic repulsion between protons becomes weaker. Elements heavier than these isotopes can yield energy by nuclear fission; lighter isotopes can yield energy by fusion.

The stability of the atoms is determined by the interactions in which its constituents are subject. The increase of Z also increases the repulsive forces of electrical nature between the protons: the nuclear stability is preserved by increasing progressively the number of neutrons, so as to balance the increase of repulsion and prevent decay of the nucleus (Figure 1.6).
The energies of stable nuclei are lower than unstable nuclei. If we report the energy associated with various nuclei in three dimensional, we find that the permanent ones form a valley, so-called valley of stability (Figure 1.7).

The unstable nuclei are located far away from the bottom of the valley; are placed at one of the two sides of the valley depending on whether their instability is generated by
an excess of neutrons or by an excess of protons. The achievement of a more stable configuration is obtained through two possible types of decay involving the emission of particles: $\alpha$ and $\beta$ decay. The instability of a nucleus can be caused not only by the overabundance of a type of nucleon, but also by an excess of energy that the nucleon possesses when it is an excited level: in this case there is a third type of decay that causes only the emission of the surplus of energy in the form of electromagnetic radiation, i.e., $\gamma$ decay.

1.3.1 Alpha decay

The alpha decay occurs in accordance with the conservation law of mass/energy through the emission from the nucleus of an element with high atomic number ($Z > 83$) of a particle, called alpha particle, consisting in two protons and two neutrons (helium nucleus). There are no $\alpha$ emitters with $A < 146$ and this can be explained by analyzing the trend of the binding energy per nucleon as a function of the mass number. A nuclear system gains in binding energy emitting $\alpha$ only if it is located beyond the maximum of the curve of figure 1.5, since in this region $\varepsilon$ increases with decreasing of the value $A$. The emission of an alpha particle from the initial nucleus leads to the reduction of mass with four units and two units of charge on the final nucleus as can be seen in the following schematic decay process:

$$\frac{4}{2}X \rightarrow \frac{4}{2-4}Y$$

(Eq. 1.15)

The process takes place only for those nuclei in which $Q = M_X - M_Y - m_\alpha > 0$. Between the initial nucleus and the final products of the decay is exchanged a fixed amount of energy $Q$, divided among the decay products in such way which is in respect with the conservation of energy and the preservation of the moment. Since this is a decay in two bodies, the energy of the particle $\alpha$ is fixed in the terms of masses of the nuclei. Because the nucleus that accompanies the particle $\alpha$ is much heavier, with a good approximation we can consider that the value $Q$ of the reaction corresponds to the kinetic energy of $\alpha$ emission. The kinetic energies involved are in the order of some MeV.

The emission of an $\alpha$ particle can leave the parent nucleus directly to the ground state of the daughter or leaving the daughter product in an excited states. The final
daughter nuclide can reach the ground state by releasing the energy through the emission of electromagnetic radiation of γ type. [Cember H., and Johnson T.E., 2009], [Das A., and Ferbel T., 2003].

Figure 1.8: schematic representation of α decay and α-γ decay of $^{227}$Th.

The energy spectrum resulting from decay is a discrete type: the various lines correspond to the transitions which can occur between the various excited levels of the daughter nucleus.

Figure 1.9: alpha spectrum associated with the decay of $^{227}$Th.
Since, $\alpha$ decay has a discrete spectrum, it is possible to conduct $\alpha$ spectrometry measurements thanks to which, on the basis of $\alpha$ emission energy, one can identify the type of atomic species subject to the process of decay.

### 1.3.2 Beta decay

The $\beta$ decay corresponds to a transition along an isobar line, which is characterized by a constant value of the mass number $A$. What occurs is the transformation of neutrons into protons in the case of nuclei with excess of neutrons (decay $\beta^-$), or on the opposite side of the valley of stability, the transformation of protons into neutrons for nuclei with excess of protons (decay $\beta^+$ or electron capture).

In the simplest beta-decay process, a free neutron decays into proton emitting an electron and an antineutrino:

$$n \rightarrow p + e^- + \bar{\nu}_e$$  \hspace{1cm} (Eq. 1.16)

The process is possible because $Q = m_n - m_p - m_e = 0.78 \text{ MeV} \ (>0)$.

The $\beta^-$ decay of a nucleus is substantially the same process, where the energy released varies depending on the binding energies of protons and neutrons in the nucleus, which determine the $Q$ value of the reaction. The transformation that is observed is of the type:

$$\frac{A}{2}X \rightarrow z + \frac{A}{2} Y + e^- + \bar{\nu}_e$$  \hspace{1cm} (Eq. 1.17)

The existence of antineutrino $\bar{\nu}_e$ as a product of $\beta^-$ decay was inferred from the observation of the energy spectrum of electrons produced in the decay. The electronic spectrum, in difference with $\alpha$ spectrum is not discrete but continuous: the kinetic energy extends from a minimum value equal to zero to a maximum value equal to $T_{\text{max}} = M_X - M_Y$.
- m_e. The conservation of energy of angular momentum and charge, respectively require that the decay is accompanied by a particle with rest mass very small, half-integer spin and neutral charge, particle corresponds precisely to the anti-neutrino. Nuclei that decay in β⁻ decay mode have mean lifetimes between 10⁻³ and 10²³ s, much longer than those of an electromagnetic decay process.

It’s quite unusual that a β transition occurs directly from a ground state of the parent nucleus to the ground state of the daughter nucleus, but in general the state of arrival is an excited state of the daughter nucleus and a β-γ decay occurs. The gamma radiation is released when the daughter nucleus in the excited state is de-energized and reaches the ground state through one or more transitions.

In figure 1.11 is shown the schematic representation decay for ¹³⁷Cs into ¹³⁷Ba: it may be affected through pure β⁻ decay to the ground state with a probability of 5.4 % or through a β-γ decay with a probability of 94.6 %.

![Figure 1.11: schematic representation decay for ¹³⁷Cs.](image)

The β⁺ decay corresponds to a transformation within a nucleus of a proton into a neutron, accompanied by the release of a positron and an electron neutrino:

\[
\frac{A}{2}X \rightarrow z\frac{A}{2}Y + e^+ + \nu_e \quad \text{(Eq. 1.18)}
\]

In this case, there is no a reaction equivalent to the decay of the free neutron described for the β⁻ decay, since the free proton is stable. This is due to the fact that the reaction \( p \rightarrow n + e^+ + \nu_e \) is prohibited by the conservation of energy, being its value Q negative. But if it is in the presence of a nucleus, could happen that the neutron is more tied with proton (\( Q = M_X - M_Y - m_e > 0 \)), and therefore in this case the reaction can take place. Since the β⁺ decay, similar to the β⁻, is a three-body decay, the energy spectrum of the positron is a continuous spectrum. The positrons have a rather short mean lifetime: they are strongly slowed down the passage on the subject, and then annihilate with electrons present in the medium in which they move. From annihilation two photons are produced, each with
energy $E_\gamma = 0.511$ MeV, equal to the rest mass of the electron. From the conservation of the momentum these photons are emitted in opposite directions.

A different physical process in competition with the $\beta^+$ decay is electron capture:

$$p + e^- = n + \nu_e \quad \text{(Eq. 1.19)}$$

Therefore electron capture corresponds to a transformation of the type:

$$\frac{A}{2}X + e^- \rightarrow Z_{-1}Y + \nu_e \quad \text{(Eq. 1.20)}$$

There is a finite probability of finding an electron in the atomic shell of the nucleus, in particular those of the lower shell, the shell K. Since an electron capture leaves a vacation in the K shell, the electrons perform a cascade to fill emitting of X-ray characteristic. The X-ray emission can be easily followed by other emissions due to the electronic cascade that is generated to fill the gap that moves progressively from the inner electronic shell in the outer ones.

![Figure 1.12: schematic representation of electron capture process.](image)

Sometimes the released energy is converted in the production of a photon: it may happen that it is transferred to a third electron, the outermost shell, which is able to reach the level of vacuum. This process is called Auger effect. The Auger electrons are monoenergetic, and usually have low energy, as they are expelled from the atomic shell for which the binding energies are weak.

1.3.3 Gamma decay

The $\gamma$ decay, unlike $\alpha$ and $\beta$ decays, doesn’t change neither the atomic mass number $A$ or the charge $Z$ of a nucleus, but involves the dissipation, in the form of electromagnetic radiation, of excess of energy of the excited nucleus. The nucleus possesses similarly to
atoms energetic levels spaced by bands of forbidden energies: when a nucleus is located in an excited state emits electromagnetic radiation in order to achieve a state of lower energy, hence more stable. Therefore the energy of gamma radiation is well-defined and equal to the difference in energy between the nuclear levels involved in the transition. The $\gamma$ decay processes are like this:

$$\frac{A}{z}X^* \rightarrow \frac{A}{z}X + \gamma$$  \hfill (Eq. 1.21)

They are qualitatively similar to atomic de-excitation, with an important difference in which the energies involved are of the order of MeV and not of eV. For gamma radiation is defined conventionally as that part of the electromagnetic spectrum that is associated with energy greater than 40 keV. As all electromagnetic waves traveling at the speed of light $c$, has a discrete energy $E$, a frequency $\nu$ and a wavelength $\lambda$, related by the following relationship:

$$E = h\nu = h\frac{c}{\lambda}$$  \hfill (Eq. 1.22)

where $h = \text{constant of Planck, equal to } 6.626 \times 10^{-34}\text{Js and } c = \text{speed of light in vacuum, equal to } 2.998 \times 10^8 \text{m/s.}$

The radiation $\gamma$ usually accompanies an alpha or beta radiation. After emitting $\alpha$ or $\beta$, the nucleus is still excited because its protons and neutrons have not yet reached a new equilibrium, and consequently the nucleus is released quickly the surplus energy by emitting electromagnetic radiation. Since the $\gamma$ radiation due to transitions between energy levels nuclear, it is obviously characteristic of the nucleus that produces it. The energy spectrum generated by a transmitter $\gamma$ is discrete because it is composed of many energetic levels that make possible the nuclear transitions.

### 1.4 Sources of radioactivity

We are exposed to environmental radiation from different sources [Klement A.W., 1982] [NCRP Report No.45, 1975]. The origin of radioactivity in the environment can be divided into two main sources: (a) natural and (b) man-made sources [Lilley J., 2001].
Mostly the naturally occurring radiation arises from terrestrial radioactive nuclides that are widely distributed in the earth’s crust and extra-terrestrial sources arising from cosmic ray [UNSCEAR 2000]. Also from human activities arise some other sources concerned with the use of radiation and radioactive materials from which releases of radionuclides into the environment may occur [Eisenbud M., and Gesell T., 1997].

1.4.1 Cosmic radiation

The result of work undertook by Victor Hess between years 1911 and 1913, has explained that there was a radiation penetrating earth’s atmosphere and originating from outside the earth, who gave the radiation the name "cosmic rays". Cosmic radiation comes from both the primary energetic protons and alpha particles of extraterrestrial origin that strike the earth’s atmosphere and the secondary particles or cosmogenic radionuclides which are continuously generated by bombardment of stable nuclides in the atmosphere from these cosmic rays. The primary cosmic rays are those who have not yet any kind of interaction with the present matter in the Earth's atmosphere. They mainly consist of protons and alpha particles, and a small part by heavier nuclei. So the primary cosmic ray composition is very heterogeneous including (about 98% in total, of which 87% consists of hydrogen, 12% of helium and 1% of heavy nuclei), with a small contribution of electrons and positrons (2%) [Bartlett D. T., 2004].

Penetration of cosmic rays depends on several factors, mention here the Earth’s magnetic field and the attenuation caused by the atmosphere, thus only a part of the cosmic radiation incident reaches the earth’s surface, irradiating all living things in continuously way, including human beings. The collision of cosmic radiation particles with atoms in the atmosphere, causes ionization and losing gradually of their energy. The process of energy loss occurs through elastic and inelastic collisions with atomic nuclei, generating a cascade of secondary radiation, as shown in figure 1.13. This secondary radiation includes neutral and charged pi mesons (π⁰ and π⁺⁻⁻), and anti-protons and anti-neutrons (p and n), heavy mesons (K) and hyperons (Y).
Figure 1.13: representation of the reactions involved in the interaction of particles of primary cosmic rays with the atmosphere, giving rise to secondary cosmic rays.

The cosmic ray flux that reaches at the Earth varies greatly depending on the geomagnetic latitude and at the sea level, consists mainly of muons, electrons and small percentage of neutrons and protons. Muons are the most penetrating components and in spectroscopy is the major source of noise due to cosmic rays. By the interaction of cosmic rays with atoms of the atmosphere through the processes of spallation or neutron capture, we have the production of cosmogenic radionuclides. The spallation is a nuclear reaction in which a nucleus splits into lighter nuclei to collide with a high-energy neutron or a charged particle. In the reaction also produces secondary neutrons which can then give rise to processes of neutron capture. Radionuclides of this nature that contribute most to natural radioactivity are: $^3$H, $^7$Be, $^{14}$C.

1.4.2 Primordial radionuclides

The primordial radionuclides which are of terrestrial origin, also called terrestrial radionuclides, have been present when the Earth formed about 4.5 billion years ago. These radionuclides have very long half-lives comparable to the age of the earth and are found around the globe in sedimentary and igneous rock. Therefore these radionuclides can migrate from rocks into soil, water and air too. The main primordial radionuclides include the series of radionuclides produced when uranium and thorium decay, as well as $^{40}$K and $^{87}$Rb.
In the Table 1.1 are shown ranges and averages concentrations of $^{238}$U, $^{232}$Th, and $^{40}$K in typical rocks and soils [Eisenbud M., and Gesell T., 1997], [IAEA TRS No.419, 2003].

Table 1.1: ranges and averages of the concentrations of $^{238}$U, $^{232}$Th, and $^{40}$K in typical rocks and soils.

<table>
<thead>
<tr>
<th>Rock type</th>
<th>Potassium-40 Total K</th>
<th>Thorium-232</th>
<th>Uranium-238</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td>%</td>
<td>Bq/kg</td>
<td>ppm</td>
</tr>
<tr>
<td>Igneous rocks</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Basalt</td>
<td>0.8</td>
<td>300</td>
<td>3.4</td>
</tr>
<tr>
<td>Mafic</td>
<td>0.3-1.1</td>
<td>70-400</td>
<td>1.6-2.7</td>
</tr>
<tr>
<td>Sialic</td>
<td>4.5</td>
<td>1100-1500</td>
<td>16.20</td>
</tr>
<tr>
<td>Granite (crystal, aver.)</td>
<td>&gt;4</td>
<td>&gt;1000</td>
<td>17</td>
</tr>
<tr>
<td>Sedimentary rocks</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Shale, sandstones</td>
<td>2.7</td>
<td>800</td>
<td>12</td>
</tr>
<tr>
<td>Clean quartz</td>
<td>&lt;1</td>
<td>&lt;300</td>
<td>&lt;2</td>
</tr>
<tr>
<td>Dirty quartz</td>
<td>2</td>
<td>400</td>
<td>3-6</td>
</tr>
<tr>
<td>Arkose</td>
<td>2.3</td>
<td>600-900</td>
<td>2</td>
</tr>
<tr>
<td>Beach sands</td>
<td>&lt;1</td>
<td>&lt;300</td>
<td>6</td>
</tr>
<tr>
<td>(limestone)</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Carbonate rocks</td>
<td>0.3</td>
<td>70</td>
<td>2</td>
</tr>
<tr>
<td>All rock (range)</td>
<td>0.3-4.5</td>
<td>700-1500</td>
<td>1.6-20</td>
</tr>
<tr>
<td>Continental crust (ave.)</td>
<td>2.8</td>
<td>850</td>
<td>10.7</td>
</tr>
<tr>
<td>Soil (ave.)</td>
<td>1.5</td>
<td>400</td>
<td>9</td>
</tr>
</tbody>
</table>

The average abundances of the continental upper crust in the world for $^{238}$U, $^{232}$Th and $^{40}$K are respectively 2.7 ppm, 10.5 ppm and 2.3% [Rudnick R.L., and Gao S., 2003].

**Potassium**

The $^{40}$K is a naturally occurring radioactive isotope of potassium and comprises a very small fraction (about 0.012%) of naturally occurring potassium. Due to the ratio between the abundance of $^{40}$K and the total abundance of potassium is fixed, the detection of gamma emission of $^{40}$K can be used to estimate the amount of potassium present in the environment. The half-life of $^{40}$K is 1.3 billion years, and it decays to $^{40}$Ca by emitting a beta particle (89% of the time) with no attendant gamma radiation and to the gas $^{40}$Ar by electron capture (11% of the time) with emission of a 1460.86 keV gamma ray.

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1 According to [IAEA TECDOC No.1363, 2003] conversion of radioelement concentration to specific activity for 1% K = 313 Bq/kg; 1 ppm U = 12.35 Bq/kg and 1 ppm Th = 4.06 Bq/kg. NOTE: These coefficients are calculated for natural isotopic abundances of 99.2745% for $^{238}$U, 100% of $^{232}$Th and 0.0118% of $^{40}$K.
Uranium

Uranium is found in nature in three different isotopes as $^{238}\text{U}$ (99.2739–99.2752%), $^{235}\text{U}$ (0.7198–0.7202%), and a very small amount of $^{234}\text{U}$ (0.0050–0.0059%). The $^{238}\text{U}$, otherwise $^{40}\text{K}$, does not reach the stability with only one decay, but gives rise to a decay chain through which reaches in the stable isotope of $^{208}\text{Pb}$. Not all of the series nuclides emit gamma radiation, therefore the detection of uranium depends on the gamma rays emitted by some decay products. The most important gamma rays for $^{238}\text{U}$ are those with energy equal to 610 keV, 1120 keV and 1740 keV originated by transition of $^{214}\text{Bi}$. The half-life of $^{238}\text{U}$ is $4.47 \cdot 10^9$ years.
Thorium

The only constituent of natural thorium is $^{232}$Th with a half-life of 1, 39 x 10$^{10}$ years. Further there are shorter half-life thorium isotopes in all three natural decay chains, like: $^{234}$Th (24.1 d half-life) and $^{230}$Th (7.54 x 10$^{4}$ y half-life) in the $^{238}$U chain; $^{228}$Th (1.9 y half-life) in the $^{232}$Th chain; and $^{231}$Th (1.06 d half-life) in the $^{235}$U chain.

Similarly to $^{238}$U gives rise to a decay chain (Figure 1.15), in which the gamma emissions most important are those produced by the transitions of $^{208}$Tl at energies of 580 keV and 2614 keV. The decay chain ends with the stable isotope of $^{208}$Pb.

Figure 1.15: uranium decay chain.
All decay products that are in the chain of uranium (Figure 1.15) and thorium (Figure 1.16) have average half-life shorter than that of the generator elements of the series, therefore if the system that contains these radionuclides is isolated, may develop the condition of secular equilibrium. However, there are factors that can lead to breakage of such condition of equilibrium, such as the relatively long life or the chemical-physical characteristics of the decay products. The disequilibrium is generated when one or more decay products within a series are completely or partially removed or added to the system. The thorium rarely leaves the condition of equilibrium, while for potassium the problem does not exist since it decays directly to a stable isotope. Instead in the case of uranium should be considered that its decay chain possesses the rings particularly weak which can lead to breakage of the condition of secular equilibrium, among which the most important is certainly the $^{222}\text{Rn}$. The $^{222}\text{Rn}$ is a radioactive element with half-life equal to 3.82 days that descends from the alpha decay from $^{226}\text{Ra}$. It is founded in the gaseous state, therefore in the presence of groundwater or splits can diffuse and reach the surface.

The concentration of radon in the air depends on many factors and can depending on the type of geographic and weather conditions (humidity, rain, snow, day-night, etc.) vary strongly. Radon is also emanating from building materials, which in general contain a certain amount of $^{226}\text{Ra}$. A significant example is given by the materials of volcanic origin,
such as tuff. The existence of $^{226}$Ra in concrete and plaster, among other very porous materials, contributes to the presence of $^{222}$Rn inside buildings. In figure 1.17 and figure 1.18 are highlighted the possible areas in which the condition of secular equilibrium is broken [IAEA, TRS No.34, 2003].

The disequilibrium in the decay chain of uranium is an important source of error: the uranium concentration is estimated by studying the decay of $^{214}$Bi, which is located in a position of decay chain away from uranium generator of the series. Uranium concentrations are therefore defined as concentrations of uranium equivalent, to specify that the values are derived in condition of a secular equilibrium. For the same reason, is evaluated the thorium in indirect mode, by using in this case also the unit of equivalent concentration, although the number of decay of thorium can be considered almost in equilibrium.

![Diagram](image_url)

*Figure 1.17: decay chain of $^{238}$U where are highlighted the possible points in which the condition of secular equilibrium is broken.*
1.4.3 Man-made radionuclides

In the environment around us there are many radioisotopes of artificial origin. They come from testing nuclear weapons, from the escape of radioactive material from nuclear power plants, waste from fission released into the environment, the use of certain industrial equipment, from research activity, etc.. The main radioactive elements of anthropogenic origin are: $^{137}\text{Cs}$, $^{239}\text{Pu}$, $^{90}\text{Sr}$, $^{60}\text{Co}$.

The $^{137}\text{Cs}$ is a radioactive isotope of cesium which is mainly generated as a product of nuclear fission. It has a half-life of 30.07 years and undergoes $\beta^-$ decays into a metastable isotope of $^{137}\text{Ba}$ in 94.6 % of cases, while the remaining 5.4% of the population decays directly on the ground state of $^{137}\text{Ba}$ through a pure decay $\beta^-$. The $^{137}\text{Cs}$ is water-soluble and very toxic. Detectable quantities of man-made radionuclides are widely distributed in the atmosphere, particularly as a result of nuclear weapons testing and the accident of Chernobyl reactor in 1986.

The concentration of $^{137}\text{Cs}$ in a determined site depends on the environmental conditions during the deposition and dynamics of sedimentation. The $^{239}\text{Pu}$ together with
\(^{235}\)U is the main fissile isotope used in the nuclear industry and has a half-life of 24.11 years. It is normally produced in nuclear reactors exposing \(^{238}\)U to a neutron flux. This is transformed into \(^{239}\)U that undergoes two rapid decays \(\beta\), transforming first in \(^{239}\)Np and subsequently in \(^{239}\)Pu. After exposure the \(^{239}\)Pu formed is mixed with a residual quantity of \(^{238}\)U and traces of other isotopes of uranium, so is subjected to a purification treatment that occurs primarily via chemical.

The \(^{60}\)Co is a radioactive isotope of cobalt, and has a half-life of 5.27 years. It is produced artificially by neutron activation of \(^{59}\)Co. The \(^{60}\)Co undergoes \(\beta^-\) decays into \(^{60}\)Ni, which emits gamma radiation with energy respectively 1.17 and 1.13 MeV. The \(^{60}\)Co is mainly used for sterilization of medical equipment, as radiation source for radiotherapy, for industrial radiography or for research purposes.

The \(^{90}\)Sr is a radioactive isotope of strontium with a half-life of 28.8 years. It undergoes \(\beta^-\) decays into \(^{90}\)Y, which in turn undergoes \(\beta^-\) decays into \(^{90}\)Zr with a half-life of 64 hours. The \(^{90}\)Sr finds many applications in medicine and industry, in particular in the field of radiotherapy surface of some types of cancer. The decay of \(^{90}\)Sr produces a lot of heat and given that \(^{90}\)Sr is cheaper than \(^{238}\)Pu, is often used as a heat source in many radioisotope and thermoelectric generators.
Chapter 2

2. Gamma-ray spectrometry and principles of gamma ray detectors

The operation of a detector is based on the interaction of photons constituting the incident radiation with the material that constitutes the detector itself. The gamma radiation interacts with matter via three main processes: the photoelectric effect, Compton scattering and pair production. Thanks to these processes, all or part of the energy possessed by the radiation is transferred to the mass of the detector and then converted into an electrical signal. The basic notions related to the interaction of electromagnetic radiation with matter that we will provide in this chapter will therefore be useful to understand the mechanisms that are at the basis of the generation of a gamma spectrum.

In addition, this chapter will briefly describe the two main types of gamma radiation detectors, i.e. the semiconductor detector and the scintillation, in particular the high-pure germanium detector (HPGe) and a sodium iodide detector activated by thallium NaI(Tl).

2.1 Photoelectric effect

The photoelectric effect is the process by which the energy of incident photon is absorbed by an atom and one of the electrons is released, creating an ion and free electron. During the photoelectric effect the incident photon energy must have higher energy than the binding energy of the inner shell electron. The electron is ejected from its shell, and acquires a kinetic energy equal to:

\[ E_e = E_\gamma - E_b = h\nu - E_b \]  \hspace{1cm} (Eq. 2.1)

Where \( E_\gamma \) is the energy carried by the photon and \( E_b \) is the energy required to remove an electron from the material, which takes the name of binding energy. The quantity \( E_b \) represents both the energy required to overcome the nuclear attraction and the energy that
is spent in the internal electronics collisions. In the case where the binding energy is minimal, the photoelectron emerges with the maximum kinetic energy $E_{\text{max}}$ that is equal to:

$$E_{\text{max}} = E_\gamma - E_b = h\nu - E_b$$  \hspace{1cm} \text{(Eq. 2.2)}$$

where $E_b$ is the work function equal to the minimum energy required in order that an electron can be extracted from the material.

Considering the case in which the maximum energy obtainable from the electron is zero can be go up to the threshold frequency $\nu_0$, below which the photoemission process cannot take place:

$$h\nu_0 = E_{b0}$$  \hspace{1cm} \text{(Eq. 2.3)}$$

The photon frequency $\nu_0$ is one of that has sufficient energy to remove the electron without giving any kinetic energy. If the frequency is less than the threshold value of the radiation, regardless of the intensity, will not have enough energy to remove the electron from the material.

The photoelectric effect leaves the atom in an excited state with an excess of energy equal to $E_b$ and a gap in the electronic orbit from which it was released the photoelectron. The gap created following the expulsion of the photoelectrons can be progressively filled by electrons from the outer orbits resulting in a characteristic X-ray emission. Alternatively, the excitation energy can be carried away by the release of other, less tightly bound electrons known as Auger electrons.

![Figure 2.1: schematic representation of the photoelectric effect and the subsequent emission of X-ray characteristic.](image)

During photoelectric process the interaction cross section ($\tau$) varies in a complex manner with $E_\gamma$ and with the $Z$ value of the absorber as follow:
\[ \tau \approx \text{const}\frac{z^n}{E^{m}} \]  \hspace{1cm} (Eq. 2.3)

where \( n \) and \( m \) are numbers that vary in the range from 3 to 5 over the gamma-ray energy region of interest. Probability of photoelectric absorption strongly depends on photon energy and atomic number of an absorber material. While the strong dependence of \( Z \) shows that a high-\( Z \) material is very effective in the absorption of photons. The strong dependence on the photon energy is the reason why the photoelectric process is significant at low energy of photons, but becomes less dominant at higher energies.

### 2.2 Compton scattering

The Compton effect is a phenomenon of scattering representable as a collision between the incident gamma-ray photon and weakly bound or free electron in the absorbing material.

\[ \lambda_f - \lambda_i = \Delta \lambda = \frac{h}{m_e c} (1 - \cos \theta) \]

**Figure 2.2**: schematic representation of the scattering between a photon and a free electron.

In the collision part of the energy of the photon is transferred to the electron by the expression:

\[ E_e = E_\gamma - E'_\gamma = E_\gamma \left[ 1 - \frac{1}{1 + \frac{E_\gamma}{m_e c^2} (1 - \cos \theta)} \right] \]  \hspace{1cm} (Eq. 2.4)

where \( m_e \) is indicated the electron mass and \( \theta \) the angle formed by the radiation deflected respect to the direction of incidence. The transfer of energy can be formulated also according to variation of the photon wavelength.
$\Delta \lambda = \lambda_c (1 - \cos \theta)$ \hspace{1cm} (Eq. 2.5)

where $\lambda_c$ is the Compton wavelength defined by:

$\lambda_c = \frac{\hbar}{mc^2}$ \hspace{1cm} (Eq. 2.6)

The energy acquired by the electron depends on the direction in which the incident photon is deflected: if $\theta = 0^\circ$, or if the direction of deviation of the photon coincides with the direction of incidence, $E_e$ is equal to 0, and therefore no energy is transferred to the electron. The photon maintains its initial energy, giving rise to a process of elastic scattering, known as Rayleigh scattering. In the case diametrically opposite, in which the photon is scattered backwards at an angle $\theta = 180^\circ$, the energy acquired by the electron is the possible maximum but less than the energy carried by the incident photon. For each possible scattering angle the percentage of energy transferred to the electron is always less than 100%.

In the Compton scattering process the probability depends strongly on the number of electrons per unit mass of the interacting material. It also depends on the incoming gamma-ray energy as function of $\frac{1}{E_{\gamma}}$ [Lilley J., 2001], [Gilmore G.R., 2008]. Compton scattering is the dominant interaction process for gamma-ray energies ranging from 0.1 to 10 MeV [Das A., and Ferbel T., 2003]. Another interaction mechanism, known as ‘pair production’ becomes more significant at the higher energy.

### 2.3 Pair production

The pair production occurs when a photon of high energy loses all its energy in the interaction with an atomic nucleus, giving rise to an electron-positron pair. This phenomenon take place if the photon energy is greater than a threshold value of 1022 keV, equal to the sum of the rest mass produced by particles, therefore equal to twice the mass of the electron.
In the pair production the energy that is transferred to the nucleus is negligible, being much more massive respect to the particles produced. Thus the total relativistic energy is given by (Eq. 2.7):

\[ E_\gamma = h\nu = E_- + E_+ = 2m_ec^2 + K_- + K_+ \quad \text{(Eq. 2.7)} \]

where \( K_- \) and \( K_+ \) indicate the kinetic energy of electron and positron respectively. The positron is produced with energy slightly higher than the electron due to the Coulomb interaction of the pair with the positively charged nucleus produces an acceleration of the positron and electron deceleration.

After creation of electron - positron pair, they can traverse the medium, losing their kinetic energy by collisions with electrons in the surrounding material through ionization, excitation and/or bremsstrahlung. When the energy of the positron reaches values close to the thermal energy, it annihilates with an electron surrounding releasing the two annihilation photons of 511 keV. The energy (expressed in keV) which is absorbed by the detector following the pair production will be equal to:

\[ E_e = E_\gamma - 1022 \text{ keV} \quad \text{(Eq. 2.8)} \]

### 2.4 The probability interaction and correction of gamma radiation

The probability that a photon interacts with matter is expressed in terms of the cross section \( \sigma \) (m²), which represents qualitatively the product of the area in which takes place
the total area for the probability that the interaction occurs the same. The cross section depends both from energy $E_\gamma$ of photons and the composition of the surrounding matter.

In the figure 2.4 is shown the trend of the cross section as a function of energy $E_\gamma$. From the graph it is clear that the photoelectric effect is the dominant phenomenon for energies below 100 keV: in this energy range are observed evident peaks in the form of the cross section, which are found in correspondence of the binding energies of the different atomic shell. Compton scattering dominates the events of medium energy, while the pair production is possible only for energies above the threshold of 1022 keV.

For the three processes of interaction between radiation and matter is possible to define the trend of the cross section as a function of photon energy $E_\gamma$ and as a function of the atomic number $Z$ of the material in which the photons pass through.

For the photoelectric effect its dependence can be approximately expressed by the following relation:

$$\sigma_{\text{photon}} \propto Z^n E_\gamma^{-3} \quad \text{with} \quad n = 4 \div 5$$

(Eq. 2.9)

The strong dependence from the atomic number indicates that a material with high value of $Z$ is very effective in the absorption of photons. The dependence of the photon energy is a reason why the photoelectric effect is dominant for low energies.

![Figure 2.4: cross section for the photoelectric effect, Compton scattering and pair production and the total cross section for an atom of Lead as a function of the photon energy.](image)
In the case of Compton scattering we have:

\[ \sigma_{\text{Compton}} \propto Z E_\gamma^{-1} \quad (\text{Eq. 2.10}) \]

The probability of Compton scattering depends on the number of electrons available as centers scatters, therefore the Compton cross section increases linearly with \( Z \). The inverse relationship with respect to \( E_\gamma \) justifies the dominance of this process at intermediate energies.

The cross section for the pair production depends on the energy \( E_\gamma \) and \( Z \) according to the following relationship:

\[ \sigma_{\text{pair}} \propto Z^2 \ln E_\gamma \quad (\text{Eq. 2.11}) \]

Although the threshold voltage is equal to 1.022 MeV, this phenomenon becomes dominant for energies above 4 MeV; consequently this process does not influence significantly the spectrum range associated to natural radioactivity.

Typically the gamma photons lose their energy through subsequent events of Compton scattering, until the scattered photon of low energy is not completely absorbed by the photoelectric effect. As a result of the interaction of the radiation with matter, the intensity of the radiation decreases with the distance from the source according to the law:

\[ I = I_0 e^{-\mu x} \quad (\text{Eq. 2.12}) \]

Where \( I_0 \) is initial intensity, \( x \) is thickness of material crossed (m), \( I \) is the intensity of the beam transmitted through the thickness \( x \) and \( \mu (m^{-1}) \) is the correction coefficient that represents the probability of interaction per unit length.

The radiation \( \alpha, \beta \) and \( \gamma \) have different penetration capacity for the same energy; this is due to the higher or lesser probability of interaction with matter. The alpha particles, being massive particles and charged, have a high probability to interact with the atoms of the crossed material; particles \( \beta \) are less massive and less charged than \( \alpha \), therefore pass a bigger distance in a certain material. The \( \gamma \) radiation is typically the most penetrating, since it is constituted by photons without mass and electrically neutral.

The range of gamma rays produced by natural radionuclides, or the average distance in the presence of matter, is equal to 700 m in the air, 0.5 m in the rocks and a few cm across the lead (Pb). Since gamma radiation is absolutely the most penetrating
component of radiation released by the decay of natural radioactive elements and anthropogenic, it is widely used in the study of environmental radioactivity.

2.5 Properties of gamma ray spectra

A spectrum of gamma radiation is a histogram that represents the distribution of the photons coming from a certain source as a function of their energy. The spectrum generated in the process of gamma decay is of discrete type, therefore ideally a gamma spectrum consists of a series of lines corresponding to transitions between different nuclear levels affecting the elements involved in the decay chain [IAEA TECDOC No.1363, 2003].

![Figure 2.5](image1)

**Figure 2.5:** gamma ray emission line spectra of potassium (IAEA-TECDOC-1363, 2003).

![Figure 2.6](image2)

**Figure 2.6:** gamma ray emission line spectra of uranium (IAEA-TECDOC-1363, 2003).
Each line spectrum shows the energy and relative intensity of gamma ray emissions in the decay series: the energy is equal to the difference of energy between two nuclear levels involved in the transition and the intensity is proportional to the branching ratio of specific channel of decay.

However, the spectra shown in Figure 2.5, Figure 2.6 and Figure 2.7, are ideal spectra that represent the energy distribution of the photons emitted by the source: in fact the monochromatic radiation emitted by the source are reduced in energy by Compton scattering in the source, in the detector, and also in matter between source and detector. Thus, the relative contribution of scattered and unscattered photons to the gamma ray fluence rate depends on the source-detector geometry and on the amount of attenuating material between the source and the detector.

The photons can give completely their energy whether through of a single photoelectric process that through multiple interactions of Compton scattering.
Figure 2.8: simulated potassium fluence rates at 300 meter height (IAEA-TECDOC-1363, 2003).

Figure 2.9: simulated uranium fluence rates at 300 meter height (IAEA-TECDOC-1363, 2003).

Figure 2.10: simulated thorium fluence rates at 300 meter height (IAEA-TECDOC-1363, 2003).
The photopeaks are superimposed to the spectrum of the photons not totally absorbed, which covers a range of energies continuously up to the maximum energy equal to that of the photon emitted from the source. When a photon transfers part of its energy to an electron of the detector material, the photon scattered with lesser energy can leave the detector making it impossible to complete absorption. Similarly may happen that the photons emitted by the source are scattered in the material which surrounds the instrument and enter in the detector following the deviation, with lower energy compared to the original photon. These events then are recorded with energy equal to the transmitted electron, less than the energy of the gamma radiation produced by the source. This type of events generates the continuous Compton within the total spectrum.

The measured spectra include also the contributions from background radiation, where the term background is referred the radiation generated by external sources compared to that under examination. Background radiation in the measurements of natural radioactivity may constitute a significant part of the total signal. There are three main sources of background: the intrinsic radioactivity of the detector, cosmic radiation and atmospheric radon.

The intrinsic radioactivity of the detector indicates the signal produced by traces of potassium, uranium and thorium present in the crystal inside to the detector, in the photomultiplier and in external equipment to the detector. Inside this category usually are included also the noise signal generated by the electronics, which affects mainly channels of low energy. The background radiation of the cosmic origin can be generated by the secondary cosmic radiation or by cosmogenic radionuclides.

The isotope of radon \(^{222}\text{Rn}\) and its decay products are the main source of background radiation. The \(^{222}\text{Rn}\) is presented in a gaseous state and can escape from rocks and soil and accumulate in the lower atmosphere. The products of its decay (\(^{214}\text{Bi}\) and \(^{214}\text{Pb}\)) attach to aerosols and dust particles and emit gamma rays on decay.

### 2.6 Principles of gamma ray detectors

In this work were carried out two different types of measures: on the one hand the laboratory measurement of concentrations of radioactive elements in soil samples, on the
other hand the acquisition of spectra in-situ. Each type of measure requires its specific type of detector chosen according to the purpose of the measure and the experimental conditions.

The detectors used are respectively the semiconductor detector and inorganic scintillation detector. The operation of both detectors is based on the interaction of photons with matter which constitutes the detector, in manner that all the part of energy possessed by the radiation is transferred to the electrons and is collected as an electrical signal. The detector response, once converted into electrical impulses, can be processed through an operation of amplification and digitization.

The incident radiation deposits a certain amount of energy $E$, which releases a charge $Q$ proportional to that is collected by the application of an electric field; the collection time varies from one detector to another because it depends on the mobility of charge carriers and the average distance they have to travel. A single photon generates a pulse of current $i(t)$, with duration equal to the time of collection $t_c$, connected to the charge released by the integral $Q$ as given in equation 2.13:

$$Q = \int_0^{t_c} i(t) \, dt$$  \hspace{1cm} (Eq. 2.13)

If the pulse shape, characteristic of the electronic chain, does not change from one event to another, the amplitude of the output signal is directly proportional to the charge $Q$ that generated it, and then it is possible to go up again the energy of the incident radiation.

In fact we have the interaction of more photons of different energy in a certain period of time, which will form a distribution of pulses, corresponding to the number of pulses of given energy $dN/dE$ in function of the same energy $E$. The total number of the pulses can be obtained by integrating the entire spectrum:

$$N_0 = \int_0^{\infty} \frac{dN}{dE} \, dE$$  \hspace{1cm} (Eq. 2.14)

The main features of the detector are sensitivity, resolution, efficiency and response time.

**Sensitivity of a detector**

The sensitivity is the ability of a detector to produce an output signal usable for a given radiation and for a given energy interval. At present there are no detectors which can
be sensitive to all radiation and all energies. For this reason the sensitivity, relatively to a particular radiation and at a given energy range, depends on:

- the cross sections of the physical processes involved in the interaction of the particle with the material of which it is made the detector.
- the size and the mass of the detector, because the volume of a detector influence the probability of interaction.
- the intrinsic noise of the detector or from the electronic part associated.
- the protective material placed around the detector, or the walls of the detector when they are needed.

The cross section and the mass determine the probability which the incident particle convert all or part of its energy in the form of ionization. If the particle is highly ionized are sufficient even materials with low density and volumes relatively small, while for the neutral particles, since the ionization is produced as a secondary process, instead the material with high density or large volumes is necessary.

The intrinsic noise and the thickness of the material with which is built the container of the active volume constitute a lower limit of the energy that can be revealed.

Resolution of a detector

The resolution of a detector is related to the ability to distinguish different values of the magnitude to be measured (energy, time, spatial coordinates,..). A good energy resolution makes it possible to separate two signals coming into energy (for example, by distinguishing the various peaks of energy from a given source), whereas a poor resolution would not allow to see these separate peaks.

The quantity used to characterize the resolution is the width at half height of the interest distribution (FWHM = Full Width at Half Maximum). In the case of a Gaussian distribution the FWHM is related to the standard deviation of relationship. The resolution is related to the FWHM as follows:

\[ R = 100 \frac{FWHM}{E_0} \]  \hspace{1cm} (Eq. 2.15)

where \( E_0 \) is the energy of photopeak.
By increasing the resolution $R$, decreases the possibility of separating two neighboring lines.

**Efficiency of a detector**

The efficiency of a detector allows relating the number of pulses registered by the apparatus of measurement with the actual number of photons emitted by the source. Not all radiation emitted by the source gives rise to a signal in the detector due to various factors such as the geometric configuration, the absorption by the material interposed between source and detector (or even self-absorption in the event of an extended source) and the time response of the instrument. The efficiency can be divided into absolute efficiency and intrinsic detection efficiency.

The absolute efficiency is determined by the ratio:

$$
\epsilon_{\text{abs}} = \frac{\text{number of events registered}}{\text{number of events emitted by the source}} 
$$

(Eq. 2.16)

and depends on the geometrical configuration of source-detector.

The intrinsic efficiency does not depend from the geometry and is determined by the energy of the incident radiation, the material constituting the detector and the thickness of the material (refers to direction of incidence of the radiation) and is determined by the ratio:

$$
\epsilon_{\text{int}} = \frac{\text{number of events registered}}{\text{number of events impinging on detector}} 
$$

(Eq. 2.17)
The absolute efficiency and the intrinsic efficiency are related by expression:

$$\varepsilon_{\text{abs}} = \varepsilon_{\text{int}} \frac{\Omega}{4\pi} \quad \text{(Eq. 2.18)}$$

where $\Omega$ is the solid angle subtended by the detector in steradians.

**Dead time**

The dead time is the minimum time interval that must separate two events in order that they be recorded as two separate pulses and therefore represents the time required for the instrument to process an event. During this time the detector can be:

- insensible; any event that comes during the dead time is lost.
- sensible; it has an overlap of events that leads to a distortion of the signal.

### 2.7 Scintillation detector features

A scintillator is a substance capable of producing visible light or ultraviolet following the release of energy by ionizing radiation. There are many scintillating materials, but not all are suitable for use as detectors. In general, a good scintillation detector must possess the following characteristics:

- a high efficiency of conversion of the energy released in the scintillation light (efficiency of scintillation);
- a characteristic time of emission very short;
- the transparency at the wavelength of their emission, in order to avoid the phenomenon of self-absorption and increase the collection of the scintillation photons;
- the linearity of response, in a range of energies sufficiently large, in order that the amount of light produced results as much as possible proportional to the deposited energy in the material;
- a spectral zone of scintillation compatible with the response of the photomultiplier;
- a good resistance to the stresses to which it may be subjected according to the specific application required;
The operation of a scintillation detector is based on excitation of the material of the detector caused by the incident radiation; the disexcited material, may give rise to a phenomenon of fluorescence or phosphorescence, respectively with the case in which the emission of light takes place in times lower or higher than $10^{-8}$ s (fast component and slow component). The scintillators are always coupled to a system of amplification of light, typically consists of a photomultiplier (PM) able to convert the weak optical signal into an electrical signal which can be subsequently processed and analyzed and which contains the information on the energy of incident radiation. It’s important that the detector-PM system is entirely isolated from ambient light. The basic elements of a scintillation detector are shown schematically in figure 2.12.

![Figure 2.12: schematic representation of a scintillation detector.](image)

The scintillating material may be either organic or inorganic. The process of fluorescence in organic scintillators originates from the transitions that occur in the structure of energy levels of a single molecule, and therefore may be observed from a particular molecular species irrespective of its state of aggregation. The scintillation light in these compounds arises from energy transitions of electrons covalent from $\pi$-molecular orbitals. There are distinct states of singlet and triplet. Associated with each electronic level there is a fine structure that corresponds to excitation of vibrational modes. The spacing between the energy levels of the various bands of molecular orbitals is the order of eV, while that between the vibrational levels is the order of tenths of eV.
The excitations of the singlet state decay at time $t \leq 10$ ps without emitting radiation (internal degradation). From the state $S_1$ is easy to decay in the ground state $S_0$ with light emission of fluorescence at any time $t = (1 \div 10) ns$.

Similarly, from the triplet state is arrived through the internal degradation in $T_1$ state and then arrive $T_0$ in a complex way with light emission of phosphorescence ($t \geq 10^{-4} s$). Between the organic scintillators remember anthracene ($C_{14}H_{10}$), stilbene ($C_{14}H_{12}$) and many plastics materials. The organic scintillators have relatively a low cost and produce extremely fast signals, with shorter duration time of some nanoseconds, which makes them particularly suitable for use as equipment of "trigger" time. The major limitation is the poor resolution energy (due to a low yield of light). Moreover, since their low value of $Z$, are not suitable for detection of photons, especially if we want to determine the energy spectrum.

The mechanism of scintillation in the inorganic scintillators is characteristic of the electronic band structure which is found in the crystals. In a pure crystal the transition of an ionizing radiation can cause transition of an electron from the valence band to the conduction band. The return of the electron to the valence band in the process of de-excitation causes the emission of a photon in the ultraviolet, which is a region of low efficiency for the photocathodes of photomultipliers. To obtain the emission of optical photons during the mechanism of de-excitation, the structure of the bands is modified by the introduction of appropriate impurities uniformly distributed within the crystal lattice. In this case we talk about scintillators doped or extrinsic. The impurities are introduced as
activators known centers give rise to energy levels, spatially localized, which are placed inside the forbidden energy band.

The ionizing radiation can transfer sufficient energy to an electron in such a way as to bring it into the conduction band: this is the classic mechanism of ionization in which the electrons and generated holes are freely moving in the lattice. It is also possible that the energy imparted to the electron is not sufficient to promote it in the conduction band, so that it remains electrostatically bound to the gap: in this case it has a weakly bound state called "exciton".

The exciton can move through the crystal but does not contribute to conduction because its total charge is zero. During its diffusion through the crystal the exciton can be captured by a center activator causing its ionization; the activator of center is hence de-excited with emission of light (Figure 2.14). The emission of light from the scintillating crystals depends strongly on temperature, as shown in figure 2.15.

![Figure 2.14: Schematic representation of the band structure of an inorganic scintillator. The centers activators introduce energy levels within the energy gap between the valence band and the conduction band.](image)

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**Figure 2.14:** schematic representation of the band structure of an inorganic scintillator. The centers activators introduce energy levels within the energy gap between the valence band and the conduction band.
The inorganic scintillators, for the same deposited energy by the incident radiation, emit much more light than those organic (almost an order of magnitude) and thus have a much better energy resolution. Another advantage, especially for the detection of photons, is their high atomic number $Z$, which makes them much more efficient.

Through light guides scintillators are coupled to a system of amplification of light, consisting of a photomultiplier thanks to which it is possible to convert the optical signal into an electrical signal.

A photomultiplier is constituted by a glass tube inside which has been practiced a vacuum, in which there is an anode and different electrodes which constitute the dynodes. The photons strike through a window input a surface called the photocathode.

From the photocathode, for the photoelectric effect are emitted electrons, called photoelectrons that are focused from one electrode toward the multiplication stage. This
stage consists of a series of electrodes each loaded to a potential higher than the previous. The first electron emitted by the photoelectric effect is accelerated due to the electric field and acquires kinetic energy. When the electron hits the first electrode of the dynode causes the secondary emission of different electrons of less energy. The structure of the system is designed in such way that each electron emitted from an electrode is accelerated and causes emission of different electrons from the successive electrode.

Thus there is a phenomenon in cascade for which a single photon that strikes the tube causes the transition of many electrons. The gain \( G \), namely the total number of electrons produced by an incident photon into a photomultiplier to \( n \) dynodes is:

\[
G = f^n \quad \text{(Eq.219)}
\]

where \( f \) is the emission factor of secondary electrons of each dynode. At the end of the sequence the electrons strike an anode and the electric impulse registered indicates the detection of the photon. The photomultipliers must be shielded magnetically, as an external magnetic field (also that terrestrial) can deflect the path of the electrons inside.

### 2.8 Semiconductor detectors features

The main characteristic of solids is the distribution of energy levels possible in energy bands separated by forbidden intervals (gap). The bands are formed due to the lattice arrangement of the atoms, in which outermost orbitals tend to overlap, since the lattice step is of the same order of magnitude of atomic dimensions. The complete band with higher energy is called valence band, the other one immediately above that can be partially full is called conduction band and is interposed between these two forbidden band. In semiconductors the band gap is small enough to ensure that when energy is supplied, the electrons of the valence band can pass into the conduction band, forming an electron-hole pair. The holes formed in the valence band behave as carriers of positive charge. In fact, when an electron fills an adjacent hole, leaves behind another and the repetition of this process generates additional positive current, typical of semiconductors. The energy needed for the creation of electron-hole pairs can be derived from the energy inside, and therefore from the temperature of detector, or from outside with the interaction of the radiation to be detected. In order to decrease the first component detectors are often
maintained in a nitrogen bath at 77 K. Inside a semiconductor the process of electrical conduction may be amplified through a process called doping. The crystals of semiconductor material may be doped by addition of small amounts of donor impurities (n-type) or acceptor (p-type), respectively, creating an excess of electrons or an excess of holes. Forming two adjacent regions with opposite doping in the same semiconductor crystal is obtained a p-n junction (Figure 2.17).

Due to the different concentrations of electrons and holes there is an initial diffusion of electrons toward the p region and the holes towards the n region. As a result, the electrons will spread to fill the holes in the p region, while the holes widespread in the region n will capture the present electrons.

Since the two regions p and n are both initially neutral, this recombination of holes and electrons causes an imbalance of charges and as an immediate consequence is to create an electric field in the junction between the two regions, in which is associated a difference potential that takes the name of contact potential. What is obtained is a dynamic equilibrium between the motion of drift and that of diffusion, and the creation of a region around the junction between the two physical materials in which the free carriers of charge are absent, called depletion region or region of charge space.

![Figure 2.17: schematic representation of a p-n junction.](image)

The depletion region constitutes the active element of a detector. The incident radiation interacts in the depletion area losing energy for ionization and producing electron-hole pairs: the free carriers of generated charge affected by the presence of the electric field, therefore it is to originate a movement of holes and electrons respectively, towards the contact to greater potential and lower potential which determines the passage of current.
The main advantage of semiconductor detectors is due to the fact that the energy to create an electron-hole pair is very low (about 1/100 of that required for the generation of a photoelectron in a scintillator) and this results in an excellent resolution.

In this work is used a high-pure germanium detector (HPGe). The Germanium has an atomic number \( Z = 32 \), energy gap between the valence band and the conduction band \( E_g = 0.67 \text{ eV} \), and \( \varepsilon \), energy needed to create an electron-hole pair, equal to 2.96 eV at a temperature of 77 K. In order to be used as detectors for gamma radiation must be kept at a very low temperature (typically 77 K) to minimize the background current caused by thermal agitation; therefore it is necessary to use a cooling system, generally constituted by a vessel Dewar in which a cistern of liquid nitrogen is kept in thermal contact with the detector.

The need to obtain materials of high purity has led to the development of technologies that, in the case of germanium, allow to reduce the concentrations of impurities up to \( 10^{10} \text{ atoms/cm}^3 \) (1 part in \( 10^{12} \)) and to obtain a depletion region of 10 mm for a reverse polarization voltage of 1000 V. The manufactured detectors with this type of high-pure germanium take the name HPGe (High-Pure Germanium detector).
Chapter 3

3. Campaign activity

In this chapter will be described the study area in which are performed the measurements of natural radioactivity. The area under consideration is the Ombrone basin located in southern Tuscany and Commune of Schio located in Region of Veneto. Also will be described the instrument used and the followed procedures to make measurements of natural radioactivity in-situ.

3.1 Study area

Ombrone basin

Hydrographic basin means the topographic area collecting waters flowing on the topsoil surface and merge into a particular hydric body which gives the name to the basin itself. Each hydrographic basin is separated from the contiguous ones by the so-called ‘watershed line’. Most of the hydrographic basins are formed from the union of more sub-basins represented by hydrographic basins of every individual affluent of the main river. The hydrographic basin differs from the hydrogeologic basin due to the latter considers the draining of superficial waters, as well as the flow of those present in the subsoil.

The regional Basin of Ombrone, located in southern Tuscany is divided into four hydrographic sub-basins: Ombrone, Albegna, Bruna and Osa and has a total area of 3.494 km². The main river of valley is Ombrone, which springs to the south-eastern of the Chianti Mountains at St. Gusmè, and after about 160 km, flows into the Tyrrhenian Sea to the south-west of Grosseto.
The hydrographic basin of F. Ombrone falls for the 3/5 in the Province of Siena and for the rest in that of Grosseto (Figure 3.1). The basin of F.Ombrone is formed by structural alignments (Dorsal) and by Neogenic-Quaternary basins.

The Dorsals are:

- The Dorsal of Medium Tuscan: formed by Montagnola-Senese and Monticiano-Roccastrada dorsal; the crop formations are composed by Carbonatic-clay-siliceous lithologies of the Palaeozoic and Mesozoic ages.
- The Rapolano Terme-M. Cetona Dorsal: formed by an inverted anticlinal ripple; by a lithological point of view, these formations are mainly those of the Falda Tuscana and those of the Ligurian Units.
- The Dorsal of Amiata Mount: formed by Murlo-Montalcino dorsal, from the volcanic building of the M. Amiata and the nucleus of Castell'Azzara; these formations belong, from the most recent, to the Neautoctonous Complex, the Ligurian Units, the Austroalpine Units and the Tuscan aquifer.

Instead, the Neogenic-Quaternary basins are:
• The Graben of Siena is confined to the east by Rapolano Terme-M. Cetona Dorsal, to the west by Montagnola Senese, to the north by Monteriggioni and M. Chianti and south by Pienza. It is filled with marine deposits of inferior - medium Pliocene made up of clay, sandy clay and sand with lenses of conglomerate.

• The Basin of Chiusdino in which neoautoctonous deposits crop out such as clays, sands and conglomerates, lacustrine formations (Pian di Feccia).

• The Radicofani Basin: is confined to the north by the threshold of Pienza (which separates it from the Graben of Siena), to the west by the Amiata dorsal, to the east by dorsal of Rapolano Terme-M. Cetona and to the south by the volcanics of the Volcanic Complex of Latera. The lithology that crops out in the prevalent way consist of clay deposits interspersed with sandy levels and pebbles.

In the **figure 3.2** is shown the geological map of the basin of Ombrone highlighted with 80 points in which were made measurements of environmental radioactivity.
Figure 3.2: geological map of the Basin of Ombrone River 1:500000 highlighted with the points in which were carried out the measurements of natural radioactivity.
Schio district

The commune of Schio (region of Veneto, northern Italy) is situated in North of Vicenza and East of the Lake of Garda. It is surrounded by the Little Dolomites (Italian Prealps) and Mount Pasubio. The area of Schio [A. Castellarin., et, al, 1968] is confined to the west by the group of M. Stivo reliefs and from the Adige Valley. To the east is confined by Plateau of the seven municipalities. Sourthen from the region of Lessinea and finally, to the north from the area of Caldonazzo lake. Its central part is characterized from the harsh reliefs constituting the mountain group of the small Dolomites, from whom they take their origin, with radially arrangement, deep engravings erosive, such as Vallarsa, Terragnolo valley, Leogra valley and other minor valleys. From A. Castellarin et al 1968, is described in details the geology of Schio. In the figure 3.3 is shown the geological map of the commune of Schio.
Figure 3.3: geological map of the commune of Schio.
3.2 A portable gamma-ray spectrometer: ZaNaI_1.0L

This work provides measures of gamma spectroscopy in situ by using a portable gamma-ray spectrometer called ZaNaI_1.0L, and on representative soil samples of the area where the measurement is made, by using high resolution gamma-ray spectrometer with HPGe detectors, called MCA_Rad system.

The ZaNaI_1.0L instrument is a portable gamma-ray spectrometer composed by a sodium iodide NaI(Tl) crystal, designed for natural and artificial radioactivity measurements for geophysical purposes. The ZaNaI_1.0L is able to acquire gamma ray spectra in static and dynamic configurations, recording both the GPS position and the main environmental parameters (temperature and pressure). The ZaNaI_1.0L can be transported easily on the shoulders of the operator and is able to measure the radioactivity of the surrounding environment for a radius that depends on the height. The percentage contribution of the received signal by the detector as a function of its height above the ground can be estimated by solving the following integral [Grasty., 1975]:

\[
\Phi = \frac{1}{4\pi} \int_{\theta=0}^{\pi/2} \int_{x=0}^{\infty} \int_{\phi=0}^{2\pi} n(x)e^{-\frac{(2x+\mu_k)sec\theta}{\cos \theta}} d\theta \, dx \, d\phi
\]  

(Eq. 3.1)

refers to the geometry shown schematically in the figure 3.3.

---

**Figure 3.4**: coordinate system used in the theoretical calculations for the variation of detector count rate with its height respect to the ground.
where:

\( h \) = height of the detector from the ground;

\( dv \) = elementary volume in a horizontal plane with thickness \( dx \), subtended by an angle \( d\theta \), at a distance \( x \) from the surface of the ground;

\( \mu \) = linear absorption coefficient of air;

\( \lambda \) = linear absorption coefficient of the ground;

\( n(x) \) = number of primary photons emitted per unit of volume per second by volume \( dv \);

As described below, the environmental radioactivity measurements were carried out by placing the detector at a height of 0.05 m, 0.5 m and 1 m. These values correspond to the height of the center of mass of the detector in the case where the backpack is placed on the ground or placed 1 m height from the ground. The graphs in figure 3.5, figure 3.6, and figure 3.7 are constructed from the equation above to a height \( h \) equal to \( h = 0.05 \text{ m} \), \( h = 0.5 \text{ m} \), and \( h = 1 \text{ m} \).

![Graph](image)

**Figure 3.5:** percentage contribution of the signal received by the detector placed at a height of 0.05 m.
For a height $h$ equal to 0.05 m, as shown in figure 3.5 the detector receives 90% of the signal from an area of 0.37 m of radius. The graph in figure 3.6 give us an information for a height $h = 0.5$ m. In this case, the detector receives 90% of the signal from an area of 3.7
m radius. While the figure 3.7 is constructed for a height h = 1 m in which the detector receives 90% of the signal from an area of 7.35 m radius.

The characteristics of the ZaNaI_1.0L instrument are:

- Measuring velocity: it is possible to realize and analyze a spectrum in situ less than 10 minutes. The system provides the possibility to make measurements both in static and dynamic modality;
- Recognition in situ of the radioactive isotopes: the ZaNaI_1.0L is able to discriminate the different natural and artificial radioactive isotopes through the spectral analysis;
- Accordance with international guidelines: the hardware and software of the devices were constructed inspired by the international guidelines of the IAEA and ANSI\(^2\).
- Facility of use: the ZaNaI_1.0L system is designed to be fully managed by a single person. The software allows realizing of measurements and making of the spectral analysis with a few clicks.
- Versatility for measurements in the extreme environments: the detector ZaNaI_1.0L has been designed to cope with stressful environmental conditions, typical of geophysical surveys. The instrument has successfully exceeded the test both prolonged activity and extreme weather conditions.

The ZaNaI_1.0L instrument (Figure 3.8) is constituted by:

- The ZaNaI_1.0L detector is mounted in a backpack. The detector used is a NaI(Tl) crystal with a cubic shape of a volume of 1 liter (10.16 cm x 10.16 cm x 10.16 cm) and energy resolution of 7.3% at 662 keV (\(^{137}\)Cs), 5.2% at 1172 and 1332 keV (\(^{60}\)Co).
- The detector is coupled with a photomultiplier (PMT) base with integrated bias supply, preamplifier and digital multichannel analyzer (MCA) that is controlled by a PC via USB connection. DigiBase incorporates all the functionalities of a chain of gamma spectrometry: high voltage, digital amplifier (DSP filter technology), ADC digital processing with trapezoidal waveform;
- Mini notebook Asus EeePC 900 with high-capacity battery (6 cells, 10400 mAh);

• GPS antenna (54 channels and 10 m accuracy) and a humidity and temperature sensor (10% accuracy) are used for positioning and weather condition measurement connected via USB to mini notebooks and managed by the software which is in the correspondence to the choice of the equipment.
• The obtained data log formats are easily imported to GIS systems and Google Earth(R).

![Figure 3.8: the ZaNaI_1.0L system configuration set-up in a backpack.](image)

*The calibration procedure of ZaNaI_1.0L instrument*

The calibration of a gamma ray detector is the procedure that makes it possible to interpret the acquired spectrum as a function of the energy associated to the events of decay and in the function of the concentration of radioactive elements present in the system under analysis. Portable instruments are usually calibrated for natural radioactivity measurements by means of standard spectra acquired at least using three concrete pads enriched in K, U and Th and a pad free of radioactivity representing the background [Hendriks et al. 2001], [Engelbrecht and Shwaiger 2008]; [Løvborg et al. 1981]. These pads are usually concrete constructions of cylindrical shape, having finite dimensions of 2–3 m in diameter and 0.3–0.5 m thick and for this reason it needed a geometrical correction due to the differences from an infinite calibration sources. In the table 3.1 are presented typical concentrations of constructed pads used to calibrate in-situ gamma-ray spectrometers [IAEA 1990].
Table 3.1: typical concentrations of constructed pads used to calibrate in-situ gamma-ray spectrometers (IAEA 1990).

<table>
<thead>
<tr>
<th>Pad</th>
<th>K (%)</th>
<th>eU (ppm)</th>
<th>eTh (ppm)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Background</td>
<td>0.5</td>
<td>1.1</td>
<td>1.4</td>
</tr>
<tr>
<td>Potassium</td>
<td>6.5</td>
<td>1.1</td>
<td>1.4</td>
</tr>
<tr>
<td>Uranium</td>
<td>0.5</td>
<td>50.0</td>
<td>1.5</td>
</tr>
<tr>
<td>Thorium</td>
<td>0.5</td>
<td>6.4</td>
<td>125.0</td>
</tr>
</tbody>
</table>

The design of an ideal pad with one radionuclide inside and with a perfect homogeneous distribution of the radioisotope in its volume is impossible, because impurities and non-homogeneities are always present. For this reason, compromises between accuracy and applicability of the method have to be weighted.

Generally the **Window Analysis Method (WAM)** is applied as spectrum analysis method to assess the concentration of potassium by studying the single peak emitted from $^{40}$K to an energy of 1460 keV, of uranium $^{238}$U through the detection of gamma rays of 1765 keV from $^{214}$Bi, while in the case of thorium $^{232}$Th are studied gamma rays of $^{208}$Tl with energy equal to 2614 keV. The IAEA recommend the following energy windows in order to evaluate the counts rate (**Table 3.2**).

**Table 3.2**: standard gamma ray energy windows recommended for natural radioelement mapping.

A generic site of calibration, it is obtained by evaluating the area under the spectrum within the energy bands previously defined, the number of counts in the windows of potassium, uranium and thorium, respectively $N_K$, $N_U$, $N_{Th}$.
The **WAM** method considers exclusively useful the signal associated to photopeaks neglecting other spectral regions which, despite being less representative, also provide a valuable complement to the analysis of the spectrum.

In our study, an alternative calibration procedure is used instead of building pads, which is based on the selection of sites characterized by a prevalent concentration of one of the natural radionuclides (Table 3.3). Since, the measurements in-situ can be affected by the specificity of the place, like the soil non-homogeneity, the ground morphology, the non-secular equilibrium in radioactive chains, the vertical distribution of $^{137}$Cs, the presence of vegetation, moisture, etc. the calibration sites should be selected according precise prescriptions:

- relatively uniform distribution of radionuclides in secular equilibrium with their products,
- plane area without any steps and large enough to be approximated as an infinite source (maximum 10 m radius) [Grasty et al 1975],
- undisturbed areas assuring relatively constant $^{137}$Cs vertical distribution,
- uniform and relatively homogeneous soil moisture content and vegetable coverage,

The concentrations reported in **Table 3.3** are the average of the measurements performed in the laboratory on the collected rock samples. This way the heterogeneity of each site is properly implemented by the errors which are dominated by the spread in the results of the collected samples in each calibration site.
Table 3.3: the average of the distribution of natural radioisotopes concentration. The errors correspond to one standard deviation. The conversion factors from Bq/kg are obtained by (IAEA-TECDOC-1363, 2003): 1%=313 Bq/kg for potassium, 1 ppm=13.25 Bq/kg for uranium and 1 ppm=4.06 Bq/kg for thorium.

<table>
<thead>
<tr>
<th>Site</th>
<th>K [S]</th>
<th>eJ [ppm]</th>
<th>eTh [ppm]</th>
</tr>
</thead>
<tbody>
<tr>
<td>CA1</td>
<td>53.9 ± 0.7</td>
<td>&lt;1.0</td>
<td>6.0 ± 0.5</td>
</tr>
<tr>
<td>CC2</td>
<td>0.06 ± 0.02</td>
<td>0.7 ± 0.3</td>
<td>0.6 ± 0.7</td>
</tr>
<tr>
<td>GC1</td>
<td>0.07 ± 0.04</td>
<td>0.27 ± 0.08</td>
<td>1.14 ± 0.13</td>
</tr>
<tr>
<td>GV1</td>
<td>4.9 ± 0.6</td>
<td>7.5 ± 1.1</td>
<td>7 ± 1</td>
</tr>
<tr>
<td>PM2</td>
<td>2.26 ± 0.05</td>
<td>2.27 ± 0.18</td>
<td>10.7 ± 0.8</td>
</tr>
<tr>
<td>RT1</td>
<td>0.10 ± 0.01</td>
<td>6.8 ± 1.1</td>
<td>1.74 ± 0.16</td>
</tr>
<tr>
<td>SM1</td>
<td>1.54 ± 0.14</td>
<td>1.6 ± 0.3</td>
<td>8.6 ± 0.6</td>
</tr>
<tr>
<td>SF2</td>
<td>2.92 ± 0.08</td>
<td>7.5 ± 0.4</td>
<td>39 ± 2</td>
</tr>
<tr>
<td>ST2</td>
<td>7.8 ± 0.9</td>
<td>36 ± 5</td>
<td>124 ± 16</td>
</tr>
</tbody>
</table>

The spectra acquired in situ are fitted by a linear combination of the fundamental spectra derived for each isotope from the calibration analysis. The events registered in each channel in the measured spectrum, N, can be expressed as:

\[ N(i) = \sum_{k=1}^{4} C_k S_k(i) + B(i) \]  \hspace{1cm} (Eq. 3.2)

where N(i) are the counts in the channel i, C_k are the concentration of the element k, S_k(i) are the associated counts to the fundamental spectrum of the element k in the channel i and B(i) are the counts in the channel i due to the intrinsic background, and the index k stays for \(^{40}\text{K}, ^{232}\text{Th}, ^{238}\text{U}, \) and \(^{137}\text{Cs}\). The activity concentrations are deduced applying the least square algorithm to rectangular matrix and minimizing the \(\chi^2\) as in the following equation:

\[ \chi^2 = \frac{1}{n-5} \sum_{i=1}^{N} \left[ \frac{(N(i) - \sum_{j=1}^{4} C_j S_j(i) - B(i))^2}{\sigma_{N(i)}} \right] \]  \hspace{1cm} (Eq. 3.3)

where the N(i) is considered Poisson distributed and n is the number of channels in the spectrum used in the analysis (867 in our case). During the calibration of the system the fundamental spectra (the S matrix) are obtained by solving Eq. (3.2) with the radionuclide concentrations (the C_k coefficients) reported in table 3.3:

\[ [S] = [C]^{-1} \times [N] \]  \hspace{1cm} (Eq. 3.4)
3.3 In-situ spectrum acquisition

Each point at which were performed the in-situ measurements, are chosen in such way to be as flat as possible and preferably at least 20 meters away from roads and man-made structures of large dimensions (houses, boundary walls, bridges, etc.). The latter arrangement is necessary in order to avoid interference during acquisition. The distance of 20 m is cautionary since, as shown in figure 3.7, 90% of the signal is received from an area of 7.35 m radius. First, in the selected site is made an acquisition to perform the energetic calibration of the instrument in order to be able to relate the acquisition channel with the energy associated with it. It is recognized the photopack of the potassium and the channel having the largest number of counts is assigned an energy equal to 1.460 MeV. If the calibration of energy has been carried out correctly even relative peaks of uranium and thorium fall inside its windows having energy values reported above in table 3.2.

Once completed the preliminary phase of instrument calibration, will proceed with the acquisition of the spectra. During the campaign activity in Ombrone basin and Schio Commune were acquired a total of 338 spectra. In particular, in order to study the influence of acquisition parameter were acquired:
• in Ombrone basin:
  - 80 spectra by placing the ZaNaI_1.0L on the ground (Figure 3.11) in correspondence with 400 soil samples collected in order to characterize the site (see below);
  - 80 spectra by placing the ZaNaI_1.0L a tripod at 1m height (Figure 3.12) in correspondence with 400 soil samples collected in order to characterize the site (see below);
  - 80 spectra by placing the ZaNaI_1.0L on a tripod at 1m height (Figure 3.12) in correspondence with 80 spectra by placing the ZaNaI_1.0L on the ground (Figure 3.11);
• in Schio Commune:
  - 89 spectra by placing the ZaNaI_1.0L on the shoulders of an operator (Figure 3.13) in correspondence with 89 spectra by placing the ZaNaI_1.0L on the ground (Figure 3.11).

Each acquisition had duration of 5 minutes.

Figure 3.11: example of acquisition of the spectrum with backpack placed on the ground.
**Figure 3.12:** example of acquisition of the spectrum with backpack placed on the tripod at 1 m height above the ground.

**Figure 3.13:** example of acquisition of the spectrum with backpack placed on the shoulders.
3.4 Soil sampling

The measurement in situ with portable instrumentation is strongly influenced by the surrounding environment and is fundamental the knowledge of the main environmental parameters and the characteristics of terrain. For this, in each chosen site was performed a measurement of radioactivity with the instrument ZaNaI_1.0L also have been taken five different soil samples to be subsequently analyzed in the laboratory. The samples were collected following a provision to "star": a central sample is extracted at the point in which it was positioned backpack and the four samples at a distance of about one meter (Figure 3.14). The soil is collected with a trowel to a maximum depth of 10 cm, avoiding to take roots, leaves or materials not representative of the soil itself. The amount removed is never less than 500 g. The bag is closed with adhesive tape and on the bag is shown the ID (Soil_X, where X is a progressive number) sample identification. In total 80 different sites were analyzed and were collected 400 samples of soil. All operations are documented by photos.

![Figure 3.14: example of the soil samples arrangement with respect to the central point in which is realized the measurement of gamma-ray spectroscopy in situ.](image)

For a better characterization of the environment in which it was carried out a measurement of natural radioactivity, in addition to recording the coordinates with the GPS antenna are measured some environmental parameters (temperature and humidity) and describes the pedological and geological characteristics of the soil in site. In fact there is a measurement procedure for soils which consists of the compilation in campaign of
geodatabase Rad_Nat.mdb in which are inserted the necessary characteristic parameters for the characterization of the soil sample. The environmental parameters and pedological to be defined in the campaign are reported in table 3.4 and table 3.5.

Table 3.4: standard pedological parameters to define in the campaign.

<table>
<thead>
<tr>
<th>Pedology parameters</th>
<th>Description</th>
</tr>
</thead>
<tbody>
<tr>
<td>Lithology</td>
<td>Both chemical and physical characteristics that define a rock in its various aspects, namely chemical and mineralogical composition, structure and texture.</td>
</tr>
<tr>
<td>Rockiness</td>
<td>Information indicating the content of rockiness in the surrounding area.</td>
</tr>
<tr>
<td>Use of soil</td>
<td>Use of soil is all that pertains to the use and management strategies of certain land cover by man.</td>
</tr>
<tr>
<td>PH of soil</td>
<td>Parameter that identifies the acidity of soil.</td>
</tr>
<tr>
<td>Colour</td>
<td>It is an important feature as it can be indicator of genetic variations and behavior.</td>
</tr>
<tr>
<td>Presence of mottles</td>
<td>Recognition of chromatic variations with respect to the predominant color of the terrigenous matrix.</td>
</tr>
<tr>
<td>Water content</td>
<td>Provides an indication of the amount of present water: according to the condition of the soil (dry, wet or saturated) varies the consistency at the same.</td>
</tr>
</tbody>
</table>
Table 3.5: Environmental parameters monitored during the in situ measurement.

<table>
<thead>
<tr>
<th>Environmental Parameters</th>
<th>Description</th>
<th>Instrumentation</th>
</tr>
</thead>
<tbody>
<tr>
<td>Configuration of acquisition</td>
<td>This parameter takes into account the geometry of the environment in which the measurement is carried out</td>
<td>NaI(Tl)</td>
</tr>
<tr>
<td>Humidity</td>
<td>Important parameter for the definition of the environmental conditions of measurement.</td>
<td>Humidity sensor</td>
</tr>
<tr>
<td>Temperature</td>
<td>Important parameter for the definition of the environmental conditions of measurement.</td>
<td>Temperature sensor</td>
</tr>
<tr>
<td>Climatic condition</td>
<td>Useful data for the definition of the environmental conditions of measurement, which vary in the presence or absence of precipitation meteorological</td>
<td>Detector campaign</td>
</tr>
<tr>
<td>Density of vegetative cover</td>
<td>Indicative information for both the in situ measurement and for the soil sampling</td>
<td>Detector campaign</td>
</tr>
<tr>
<td>Tipology of vegetative cover</td>
<td>Information of interest for a more detailed analysis of the environment, both for the in situ measurement and for the soil sample</td>
<td>Detector campaign</td>
</tr>
<tr>
<td>Number of soil sampling</td>
<td>Necessary parameter for ascending to the environmental conditions of measurement</td>
<td>Detector campaign</td>
</tr>
<tr>
<td>Measurement distance from the sampling point in the soil</td>
<td>Important parameter to be used for the study of the environmental conditions</td>
<td>Detector campaign</td>
</tr>
<tr>
<td>Spatial coordinates of the acquisition point</td>
<td>Information fundamental to be able to ascend the point of sampling and measuring</td>
<td>GPS 54 channels</td>
</tr>
</tbody>
</table>
Chapter 4

4. Laboratory activity

This chapter will describe the procedure for the preparation of soil samples to be analyzed with the MCA_Rad system.

4.1 Preparation of samples

Each of 400 samples was crushed, homogenized (in less than 2 mm fine grain size) and constant weight dried at 110°C temperature (Figure 4.1). Then, the samples were sealed in cylindrical polyethylene plastic containers (7.5 cm of diameter and 4.5 cm height) of 180 cm$^3$ effective volume. The soil samples were subsequently labeled with ID of the sample and after that were weighed. After, each sealed container is left undisturbed for at least four weeks before spectrum acquisition in order to establish the equilibrium between $^{226}$Ra and its short lived progenies ($^{222}$Rn).

![Figure 4.1: ventilated oven used to dry the soil samples.](image)

Each box is also assigned a bar code (Figure 4.2), necessary for their automatic identification. All data obtained from this procedure are stored in the Rad_Nat.mdb geodatabase. The material not used is kept sealed in the original plastic bags.
4.2 High resolution gamma-ray spectrometry using the MCA_Rad system

The MCA_Rad system (Figure 4.3) introduces an innovative configuration of a laboratory high-resolution gamma-ray spectrometer featured with a complete automation measurement process, which can conduct measurements on each type of material (solid, liquid or gaseous) in less than 1 hour.

The activity concentrations of $^{40}\text{K}$, $^{238}\text{U}$ and $^{232}\text{Th}$ were measured by using the MCA_Rad system, described in details in [Xhixha, G., et, al, 2012]. The MCA_Rad system is made of two 60 % relative efficiency coaxial p-type HPGe gamma-ray detectors, which possess an energy resolution of about 1.9 keV at 1332.5 keV ($^{60}\text{Co}$). The detectors are accurately shielded and positioned facing each other 5 cm apart. Both detectors are controlled by individual integrated gamma spectrometers for the digital signal processing. A new mechanical coolers, allows to simplify the management of the system. In table 4.1 the features of two detectors used for design of MCA_Rad system are presented.
Thanks to a shielding construction of lead with dimensions 20 cm × 25 cm × 20 cm and copper with dimensions 10 cm × 15 cm × 10 cm (Figure 4.4) is reached a background reduction of two order of magnitude with respect to laboratory radioactivity (Figure 4.5).
The calibration of absolute full energy peak efficiency was determined through certified standard point sources ($^{152}$Eu and $^{56}$Co). The absolute efficiency calibration is validated through the measurement of certified reference materials RGK_1, RGU_1 and RGTh_1 [IAEA, 1987], showing an overall uncertainty to be then 5%. The absolute efficiency of the standard point source $\varepsilon_p(E)$ in the energetic range from 200 to 3000 keV is obtained by fitting the function described by Knoll 1999 (Figure 4.6).
Figure 4.6: absolute efficiency curve for the MCA_Rad system obtained by fitting the corrected values for coincidence summing with equation 2.16. Apparent efficiencies of $^{152}$Eu (blue triangles) and $^{56}$Co (green squares) are also presented.

$$\epsilon(\%) = \left(1.38 \frac{E}{E_0}\right)^{1.41} + 22.97 \exp \left(-5.43 \frac{E}{E_0}\right) + 6.61 \exp \left(-0.44 \frac{E}{E_0}\right) \quad \text{(Eq. 4.1)}$$

where $E$ is the energy of gamma ray and $E_0$ is introduced to make the argument of the logarithm dimensionless. The correction due to coincidence summing is studied by following the method described in [De Felice P., et al., 2000] and obtained as a relationship between the single total efficiency ($\epsilon_t$), the single apparent absolute photopeak efficiency ($\epsilon_{app}$) and isotope decay data.

The spectra obtained are corrected for: (1) differences between the geometry of the point sources and the sample shape, CG and (2) self-attenuation, $C_{SA}$, of gamma-rays within the sample volume.

The geometrical factor for MCA_Rad system is deduced from a set of measurements using $^{56}$Co and $^{57}$Co point sources placed at different radial distances from the detector axis on three different planes.

$$C_G (E_i) = \frac{\tilde{B}_E}{\tilde{B}_{ref}} \approx \frac{\sum_{j=1}^{N} R_j(E_i)}{R_{ref}(E_i)} \quad \text{(Eq. 4.2)}$$
where \( R_x(E_i) \) is the net count rate in the standard spectrum collected in different positions \((j)\) and \( R_{\text{ref}}(E_i) \) is the net count rate in the standard spectrum collected in the reference positions (center). The geometrical correction factor was then fitted as a function of energy and considered independently in order correct the efficiency curve of the standard point source.

\[
C_G(E) = \sum_{i=0}^{n} a_i \left( \frac{E}{E_0} \right)^i
\]  \hspace{1cm} (Eq. 4.3)

where \( n = 4 \) and \( a_i \) are the fitting coefficients \((a_1 = 0.8678, a_2 = 0.1098, a_3 = -0.0541, a_4 = 0.0077)\).

The self-absorption correction factor was calculated following a simple approach discussed by Bolivar et al. 1997 and Cutshall et al. 1983:

\[
C_{\text{SA}} = \frac{1 - e^{-(\mu_s \rho_s - \mu_{\text{ref}} \rho_{\text{ref}}) t}}{(\mu_s \rho_s - \mu_{\text{ref}} \rho_{\text{ref}}) t}
\]  \hspace{1cm} (Eq. 4.4)

where \( \mu_s \) and \( \mu_{\text{ref}} \), are the densities of test sample and reference material, \( \rho_s \) and \( \rho_{\text{ref}} \), are the densities of sample and reference material and \( t \) is the sample thickness.

The self-attenuation correction factor was then given as a function of energy and considered independently in order correct the efficiency curve of the standard point source.

\[
C_{\text{SA}} = (E, \rho_s) = \exp[(1.2609 - 0.2547 \ln(E) + 0.0134 \ln(E)^2) \rho_s]
\]  \hspace{1cm} (Eq. 4.5)

The soil samples, contained in cylindrical containers of polycarbonate, are placed on a loader tilted (Figure 4.7) at the end of which is positioned a reader of bar code and a piston which introduces the sample into the measuring chamber, between the two faces of the detector, through an opening rectangular of \( 5 \times 10 \text{ cm}^2 \). A fully automation system enables up to 24 samples to be measured without any human attendance. An acquisition unit multichannel data managed by software developed in the Windows ambient controls the procedures of measurement and the change of the sample. The electronics associated with the detector and measurement of gamma-ray spectroscopy is managed by the Maestro software installed on the PC that controls the entire MCA_Rad system.
The activity concentration of $^{40}$K was determined from gamma energy emissions at 1460 keV. While the specific activity concentrations of $^{238}$U and $^{232}$Th were instead determined, assuming the secular equilibrium in their respective decay chains, from gamma energy emissions at 609 keV ($^{214}$Bi), 351.9 keV ($^{214}$Pb) for $^{238}$U and 911.2 keV ($^{228}$Ac), 238.6 keV ($^{212}$Pb), 583.2 keV ($^{208}$Tl) for $^{232}$Th. Samples were firstly measured for 1 hour typical counting time, and repeating the measurements for longer acquisition times until statistical errors become less than 20% (when reasonably possible).

The specific activity calculated using the formula:

$$A (Bq/kg) = \frac{R}{\epsilon I_{\gamma} m} C_{SA} C_G C_{CS}^r$$

(Eq. 4.6)

where $R$ is the measured count rate (background corrected), $\epsilon$ is the absolute efficiency, $I_{\gamma}$ is the gamma line intensity, $m$ is the mass of the sample, $C_{SA}$ is the correction factors for self-absorption, $C_G$ is the geometrical correction factor and $C_{CS}^r$ is the coincidence summing correction factor.
Chapter 5

5. Discussion of obtained results

In the first part of this chapter will be discussed briefly the methodology used for the analysis of spectra. In the second one the analysis of the obtained results in particular has focused on the study of four different types of correlation: correlation between in-situ acquisition on ground and laboratory measurements, correlation between in-situ acquisition on tripod and laboratory measurements, correlation between in-situ acquisition on ground and on tripod and correlation between in-situ acquisition on ground and on operator shoulder.

5.1 Spectrum analysis

A spectrum gamma acquired in natural environment is mainly composed by a photopeak associated with the decay of $^{40}$K, more than a dozen of emission photopeaks associated with the decay of $^{238}$U, and a similar number of photopeaks for the $^{232}$Th. However only some of them are distinguishable and well-formed above the spectrum of background, and are based on the principle that identifies the photopeaks relatively to which make a count of events.

As described above the Full Spectrum Analysis (FSA) reconstruct the measured spectrum composing linearly the entire fundamental spectra of potassium, uranium and thorium. The FSA method therefore uses all counts corresponding to the energy range from 300 to 2900 keV and not only those obtained in the standard energy windows (Figure 5.1).
Therefore with this method determines the shape of the spectrum from the superimposition of the different spectral components (Figure 5.1). The weight of each component is given from the isotopic abundance, that as a result of a chi-square minimization process (Eq. 3.3).

\[ [C] = [S]^{-1} [N] \]  

(Eq. 5.1)

5.2 Soil sample data: statistical analysis

From the analysis of the spectra are obtained the rate of net counts for the $^{40}$K, $^{238}$U and $^{232}$Th. Taking into account the density of the sample, self-absorption and efficiency of the system, are obtained the specific activities in Bq/kg. By using the conversion factors given in table 5.1 are obtained the abundances of K (%), eU (ppm) and eTh (ppm).

Table 5.1: conversion factors between specific activity and concentrations of K, eU and eTh.

<table>
<thead>
<tr>
<th>Concentration in the rock</th>
<th>Specific Activity (Bq/kg)</th>
<th>Conversion Factor</th>
</tr>
</thead>
<tbody>
<tr>
<td>1 % K in the rock</td>
<td>313</td>
<td>$^{40}$K</td>
</tr>
<tr>
<td>1 ppm U in the rock</td>
<td>12.35</td>
<td>$^{238}$U or $^{230}$Ra</td>
</tr>
<tr>
<td>1 ppm Th in the rock</td>
<td>4.06</td>
<td>$^{232}$Th</td>
</tr>
</tbody>
</table>

All measured parameters and spectra obtained during this procedure are recorded in the Rad_Nat.mdb geodatabase.

It should be noted that there are various factors that influence the precision of the analysis; in particular, in addition to the intrinsic radioactivity of the detector and the presence of cosmic rays, must be taken into account the statistical nature of the decay process (therefore related to the acquisition time), the variability of the radiation due to atmospheric radon, the environmental characteristics, and the geometry of the source.

In order to carry out a study on the correlation between measures of environmental radioactivity and radioactivity measurements in soil samples must have the same number of parameters to compare. It is therefore necessary to obtain a single value of the abundances of $^{40}\text{K}$, $^{238}\text{U}$ and $^{232}\text{Th}$ for each group of five samples of soil belonging to the same site. In many geological and environmental situations, the data distributions tend to be positively asymmetric.

Often, if it is proceed the logarithms of the original values, where is obtained a series of normally distributed data. If the variable $X$, has log-normal distribution, the variable $Y$ has a normal distribution such that:

$$Y = \ln(X) \quad \text{(Eq. 5.2)}$$

The density function of the lognormal variable $X$ is defined by the following equation:

$$f(X) = \frac{1}{X\sigma\sqrt{2\pi}} e^{-\frac{(\ln X - \mu)^2}{2\sigma^2}}, \quad X > 0 \quad \text{(Eq. 5.4)}$$

where $\mu$ and $\sigma$ is the median of the distribution of the logarithms of the $X$ values and the $\sigma$ is their standard deviation. If the mean, median and variance of $X$ are denoted respectively by $\mu$, and $\sigma^2$ the relationships between the parameters expressed in the original scale and in that logarithmic one are summarized in Table 5.2.

<table>
<thead>
<tr>
<th>Parameters</th>
<th>Log-normal distribution</th>
</tr>
</thead>
<tbody>
<tr>
<td>Mode</td>
<td>$e^{\mu - \sigma^2}$</td>
</tr>
<tr>
<td>Mean</td>
<td>$e^{\mu + \sigma^2/2}$</td>
</tr>
<tr>
<td>Variance</td>
<td>$(e^{\sigma^2} - 1)e^{2\sigma^2/2} + \sigma^2$</td>
</tr>
</tbody>
</table>

Table 5.2: relationships between the parameters of the distribution in the original scale and in that logarithmic one.
As an example, in **figure 5.5** is shown the frequency histogram of the abundances of $^{40}$K obtained for the 400 soil sampling, where it can be seen a slight positive asymmetry of the distribution. The histogram has been fitted with a lognormal function that can represent quite well the positive asymmetry of the distribution [Ahrens L.H., 1954]. Consider the relationships shown in **Table 5.2** were obtained average values, with the relative errors, the abundances of $^{40}$K, $^{238}$U and $^{232}$Th for each group of five samples of soil belonging to the same site.

![Histogram](image.png)

**Figure 5.2**: frequency histogram of the abundances of $^{40}$K obtained for the 400 soil sampling fitted with a log-normal function.

### 5.3 Study of performances of in-situ gamma-ray measurements

During in-situ measurements in general several sources of background radiation and environmental factors can interfere the radiation emitted from potassium, uranium and thorium present in the soil and rocks.

- Since the cosmic rays are gamma rays with high energy interact with particles in the upper atmosphere causing a spectrum of background gamma rays in the detector. Furthermore, the choice of the site can influence to this contribution due to the effect of doubling the cosmic radiation of every 1.5 - 2 km of height above sea level.
• The atmospheric radon (\(^{222}\text{Rn}\)) gas, can cause significant interference in the uranium signal, particularly when atmospheric inversion traps them close to the ground [Darnley and Grasty, 1970], [Grasty, 1979b].

• Rain and soil moisture increases the attenuation of gamma rays and cause the fluence rate of potassium and thorium to be decreased, generally by 10% for every 10% increase in soil moisture [Kogan et al., 1969]. In the case of uranium the effect is more complicated because radon escapes more readily from drier soils [Megumi and Mamuro, 1973]. In wetter soils the lower escape rate can result in a build-up of radon in the near surface and an apparent increase in uranium content [Tanner, 1964], [Lovborg, 1984]; [Grasty, 1997]. Also during rainfall the uranium signal can increase rapidly, due to daughter products of radon disintegration, attach to airborne aerosols and dust and can precipitate with the rain drops [Charbonneau and Darnley, 1970].

• The vegetative cover, having dense composition, can cause some shielding of radiation and may also contribute to the gamma ray flux through radioisotopes within the biomass. Evident is in the forested areas, tree trunks can collimate the flux of gamma rays [Kogan et al., 1969], [Travassos and Pires, 1994].

    In this study we have neglected the cosmic radiation variation since the sites are chosen at similar heights above sea-level. The effect of rainfall on soil moisture and gamma-ray attenuation is also neglected since all in-situ measurements are performed always at least 1 week after rainfalls. Instead the radon influences on uranium signal it is discussed in our results together with the attenuation effect of vegetative cover.

    In particular, we have studied the performances of in-situ measurements by studying the correlation between in-situ measurements on-ground/tripod and laboratory measurements on sample. Furthermore we have studied the effect of the morphology of the site and the attenuation factor of 1 m air by observing the correlation between in-situ measurements obtained by placing the ZaNaI_1.0L on ground and on tripod. Also the attenuation effect of the presence of the operator during in-situ measurements obtained by placing the detector on-shoulder was studied.
5.3.1 Correlation between in-situ acquisition on ground and laboratory measurements

For studying the correlation between the measurements of environmental radioactivity measured in situ with the instrument ZaNaI_1.0L and the results obtained from the analysis of soil samples, has been compared the relative abundances of $^{40}$K, $^{238}$U and $^{232}$Th obtained by MCA_Rad system with those obtained by ZaNaI_1.0L on the ground (Table 5.3). For each distribution was calculated the weighted linear regression for errors using a MATLAB code by imposing its passage through the origin of the coordinate system.

![Graph depicting the correlation between abundance of K measured using HPGe detectors and abundance of K measured in-situ using ZaNaI_1.0L placed on the ground.](image)

*Figure 5.3: the correlation between the abundance of K measured on sample using HPGe detectors and the abundance of K measured in-situ using ZaNaI_1.0L placed on the ground is described by the relationship*

\[ K_{\text{ZaNaI}} = (1.16 \pm 0.05) \times K_{\text{HPGe}} \text{ with } r^2 = 0.90. \]
Figure 5.4: the correlation between the abundance of eU measured on sample using HPGe detectors and the abundance of eU measured in-situ using ZaNaI_1.0L placed on the ground is described by the relationship

$$eU_{ZaNaI} = (0.85 \pm 0.11) \cdot eU_{HPGe} \text{ with } r^2 = 0.64.$$ 

Figure 5.5: the correlation between the abundance of eTh measured on sample using HPGe detectors and the abundance of eTh measured in-situ using ZaNaI_1.0L placed on the ground is described by the relationship

$$eTh_{ZaNaI} = (0.97 \pm 0.12) \cdot eTh_{HPGe} \text{ with } r^2 = 0.80.$$ 

The bar of errors relative to the values obtained with the MCA_Rad represents the standard deviation of the log-normal distribution of the 5 samples related to the measured site. The error bar relative to the values obtained with the ZaNaI_1.0L is the instrumental error. In table 5.3 are summarized the parameters of correlation between in-situ and laboratory measurements.
Table 5.3: the correlation parameters obtained for in-situ measurements on ground and laboratory measurements.

<table>
<thead>
<tr>
<th>Isotopes</th>
<th>$a \pm \sigma_a$</th>
<th>$r^2$</th>
</tr>
</thead>
<tbody>
<tr>
<td>K [%]</td>
<td>1.16 ± 0.05</td>
<td>0.90</td>
</tr>
<tr>
<td>U [mg/Kg]</td>
<td>0.85 ± 0.11</td>
<td>0.64</td>
</tr>
<tr>
<td>Th [mg/Kg]</td>
<td>0.97 ± 0.12</td>
<td>0.80</td>
</tr>
</tbody>
</table>

From these results it can be seen that there is good correlation between in-situ and laboratory measurements. In particular, the linear regression coefficient $r^2$ obtained for K and Th is very close to unity. While, the linear regression coefficient obtained for U show a relatively good correlation when as it can be expected the data are more dispersed due to the effect of atmospheric radon (no correction for atmospheric radon are applied to in-situ measurements). For Th the data are comparable within 1σ, while for U within 1.5σ and for K for more than 3σ. However, the final relative uncertainties for K, U and Th are less than about 20%, respectively 13%, 19% and 12%.

5.3.2 Correlation between in-situ acquisition on tripod and laboratory measurements

The previous study was repeated by comparing in-situ measurements using ZaNaI_1.0L placed on tripod and the results obtained from soil sample measurements using the MCA_Rad system.
Figure 5.6: the correlation between the abundance of K measured on sample using HPGe detectors and the abundance of K measured in-situ using ZaNaI_1.0L placed on the tripod is described by the relationship

$$K_{ZaNaI} = (1.11 \pm 0.05) K_{HPGe} \text{ with } r^2 = 0.88.$$ 

Figure 5.7: the correlation between the abundance of eU measured on sample using HPGe detectors and the abundance of eU measured in-situ using ZaNaI_1.0L placed on the tripod is described by the relationship

$$eU_{ZaNaI} = (0.75 \pm 0.10) eU_{HPGe} \text{ with } r^2 = 0.66.$$
Figure 5.8: the correlation between the abundance of eTh measured on sample using HPGe detectors and the abundance of eTh measured in-situ using ZaNaI_1.0L placed on the tripod is described by the relationship

\[ e\text{Th}_{\text{ZaNaI}} = (0.92 \pm 0.11) e\text{Th}_{\text{HPGe}} \text{ with } r^2 = 0.79. \]

In table 5.4 are summarized the parameters of correlation between in-situ and laboratory measurements.

<table>
<thead>
<tr>
<th>Isotopes</th>
<th>( a \pm \sigma_a )</th>
<th>( r^2 )</th>
</tr>
</thead>
<tbody>
<tr>
<td>K [%]</td>
<td>1.11 ( \pm 0.05 )</td>
<td>0.88</td>
</tr>
<tr>
<td>U [mg/Kg]</td>
<td>0.75 ( \pm 0.10 )</td>
<td>0.66</td>
</tr>
<tr>
<td>Th [mg/Kg]</td>
<td>0.92 ( \pm 0.11 )</td>
<td>0.79</td>
</tr>
</tbody>
</table>

From these results it can be seen that there is good correlation between in-situ and laboratory measurements. The linear regression coefficient shows similar results as those obtained for in-situ measurements placing the detector on ground. For Th the data are comparable within 1\( \sigma \), while for U within 2.5\( \sigma \) and for K for more than 2\( \sigma \). However, the final relative uncertainties for K, U and Th are less than about 35%, respectively 10%, 33% and 12%. The increase of discrepancy between the two data sets can be attributed to the attenuation of 1m air for in-situ measurement performed by placing the detector on tripod.
The attenuation due to 1 m air can be calculated as the difference between ground and tripod in-situ measurements compared with laboratory measurements, and are 5 ± 0.3 %, 10 ± 1.9 % and 5 ± 0.9 % respectively for K, U and Th.

5.3.3 Correlation between in-situ acquisition on ground and on tripod

To quantify the correction of signal mainly caused by the presence of about 1 m air between the detector and the source (earth surface) is done also by comparing in-situ measurements obtained by placing the ZaNaI_1.0L on ground and on tripod. The definition of these correction, allows to refer the measured abundances of $^{40}$K, $^{238}$U and $^{232}$Th to soil concentration In addition, also the analysis of $^{137}$Cs was included.

![Figure 5.9](image.png)

**Figure 5.9:** the correlation between the abundance of K measured by placing the ZaNaI_1.0L on ground and on tripod is described by the relationship $K_{\text{ground}} = (0.93 \pm 0.03) K_{\text{tripod}}$ with $r^2 = 0.98$. 
Figure 5.10: the correlation between the abundance of eU measured by placing the ZaNaI_1.0L on ground and on tripod is described by the relationship $e_{U_{\text{ground}}} = (0.87 \pm 0.03) e_{U_{\text{tripod}}} + (0.31 \pm 0.14)$ with $r^2 = 0.73$.

Figure 5.11: the correlation between the abundance of eTh measured by placing the ZaNaI_1.0L on ground and on tripod is described by the relationship $e_{\text{Th}_{\text{ground}}} = (0.94 \pm 0.06) e_{\text{Th}_{\text{tripod}}}$ with $r^2 = 0.96$. 
Figure 5.12: the correlation between the abundance of $^{137}$Cs measured by placing the ZaNaI$_{1.0}$L on ground and on tripod is described by the relationship $^{137}$Cs$_{\text{ground}} = (0.81 \pm 0.02) \cdot ^{137}$Cs$_{\text{tripod}}$ with $r^2 = 0.95$.

The bar of the errors relative to the obtained abundances with the ZaNaI$_{1.0}$L both on ground and the height 1 m represents the instrumental error, and thus results to be the same for both measurements.

In Table 5.5 are summarized the parameters of correlation between in-situ and laboratory measurements.

Table 5.5: correlation parameters between in-situ measurements on ground and on tripod.

<table>
<thead>
<tr>
<th>Isotopes</th>
<th>$a \pm \sigma_a$</th>
<th>$b \pm \sigma_b$</th>
<th>$r^2$</th>
</tr>
</thead>
<tbody>
<tr>
<td>K [%]</td>
<td>0.93 ± 0.03</td>
<td>-</td>
<td>0.98</td>
</tr>
<tr>
<td>U [mg/Kg]</td>
<td>0.87 ± 0.06</td>
<td>0.31 ± 0.14</td>
<td>0.73</td>
</tr>
<tr>
<td>Th [mg/Kg]</td>
<td>0.94 ± 0.06</td>
<td>-</td>
<td>0.96</td>
</tr>
<tr>
<td>$^{137}$Cs [cps]</td>
<td>0.81 ± 0.02</td>
<td>-</td>
<td>0.95</td>
</tr>
</tbody>
</table>

From these results it can be seen that there is very good correlation between in-situ measurements. Since the detector at 1 m height receives the signal from a bigger area with respect to the same positioned on the ground (Figure 3.5, Figure 3.6, and Figure 3.7), the close to unity linear regression coefficient are an evidence of the homogeneity of the
selected sites. The deviation between the angular coefficients and the unity value quantifies the correction of the signal due to the attenuation effect of 1 m air, obtaining for $^{40}$K, $^{238}$U, $^{232}$Th and $^{137}$Cs respectively 7 ± 0.3%, 13 ± 0.9 %, 6 ± 0.4% and 19 ± 0.5%. Compared with the results described in section 5.3.2 the results are comparable within 1σ for Th and U while more than 3σ away for K. For $^{137}$Cs the attenuation is higher due to the fact that it emits a gamma ray with relative lower energy (662 keV).

5.3.4 Correlation between in-situ acquisition on ground and on operator shoulder

During dynamic spectrum acquisition, by holding the detector on the operator shoulders, it is important to evaluate the interference of the operator. For this reason, a comparison between the abundances obtained with the ZaNaI_1.0L placed on the ground and held in the shoulder of an operator is performed.

![Figure 5.13](image)

Figure 5.13: the correlation between the abundance of K measured by placing the ZaNaI_1.0L on ground and on shoulder is described by the relationship $K_{\text{ground}} = (0.82 \pm 0.01) K_{\text{shoulder}} + (0.08 \pm 0.01)$ with $r^2 = 0.97$. 
Figure 5.14: the correlation between the abundance of eU measured by placing the ZaNal_1.0L on ground and on shoulder is described by the relationship $eU_{\text{ground}} = (0.84 \pm 0.01) eU_{\text{shoulder}} + (0.13 \pm 0.03)$ with $r^2 = 0.98$.

Figure 5.15: the correlation between the abundance of eTh measured by placing the ZaNal_1.0L on ground and on shoulder is described by the relationship $eTh_{\text{ground}} = (0.83 \pm 0.02) eTh_{\text{shoulder}}$ with $r^2 = 0.97$. 
Figure 5.16: the correlation between the abundance of $^{137}\text{Cs}$ measured by placing the ZaNaI 1.0L on ground and on shoulder is described by the relationship $^{137}\text{Cs}_{\text{ground}} = (0.77 \pm 0.01) \times ^{137}\text{Cs}_{\text{shoulder}}$ with $r^2 = 0.95$.

In Table 5.6 are summarized the parameters of correlation between in-situ and laboratory measurements.

**Table 5.6:** correlation parameters between in-situ measurements on ground and on shoulder.

<table>
<thead>
<tr>
<th>Isotopes</th>
<th>$a \pm \sigma_a$</th>
<th>$b \pm \sigma_b$</th>
<th>$r^2$</th>
</tr>
</thead>
<tbody>
<tr>
<td>$^40\text{K}$ [%]</td>
<td>0.82 ± 0.01</td>
<td>0.08 ± 0.01</td>
<td>0.97</td>
</tr>
<tr>
<td>$^238\text{U}$ [mg/Kg]</td>
<td>0.84 ± 0.01</td>
<td>0.13 ± 0.03</td>
<td>0.98</td>
</tr>
<tr>
<td>$^232\text{Th}$ [mg/Kg]</td>
<td>0.83 ± 0.02</td>
<td>-</td>
<td>0.97</td>
</tr>
<tr>
<td>$^{137}\text{Cs}$ (cps)</td>
<td>0.77 ± 0.01</td>
<td>-</td>
<td>0.95</td>
</tr>
</tbody>
</table>

From these results it can be seen that there is very good correlation between in-situ measurements. The deviation between the angular coefficients and the unity value quantifies the correction of the signal due to the presence of an operator, obtaining for $^40\text{K}$, $^{238}\text{U}$, $^{232}\text{Th}$ and $^{137}\text{Cs}$ respectively $18 \pm 0.2\%$, $16 \pm 0.2\%$, $17 \pm 0.4\%$ and $23 \pm 0.3\%$. 
5.3.5 *Interference of vegetative cover for-situ acquisition on ground and on operator shoulder*

During the in-situ measurements may by differences between sites due to the presence or not of the vegetative cover. We make and attempt to evaluate such a presence by classifying the sites into two classes of vegetative coverage from 0-50% and 50-100%. Therefore, we compared the results obtained by in-situ measurements by placing the detector on ground and holding it on shoulders. In the case of $^{40}$K and $^{238}$U this influence is found to be negligible within the reported uncertainty, while for $^{232}$Th and $^{137}$Cs it is observed a slight interference due to the presence of the vegetative cover.

![Figure 5.17](image)

*Figure 5.17:* for 0-50% vegetative cover case: the correlation between the abundance of Th measured by placing the ZaNaI_1.0L on ground and on shoulder is described by the relationship

$$\text{Th}_{\text{ground}} = (0.83 \pm 0.02) \cdot \text{Th}_{\text{shoulder}}$$

with $r^2 = 0.99$. 

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Figure 5.18: for 50-100% vegetative cover case: the correlation between the abundance of Th measured by placing the ZaNaI_1.0L on ground and on shoulder is described by the relationship

\[ \text{Th}_{\text{ground}} = (0.85 \pm 0.02) \times \text{Th}_{\text{shoulder}} \text{ with } r^2 = 0.97. \]

Figure 5.19: for 0-50% vegetative cover case: the correlation between the abundance of $^{137}$Cs measured by placing the ZaNaI_1.0L on ground and on shoulder is described by the relationship

\[ ^{137}\text{Cs}_{\text{ground}} = (0.72 \pm 0.02) \times ^{137}\text{Cs}_{\text{shoulder}} \text{ with } r^2 = 0.97. \]
Figure 5.20: for 50-100% vegetative cover case: the correlation between the abundance of 137Cs measured by placing the ZnNaI_1.0L on ground and on shoulder is described by the relationship

\[ 137\text{Cs}_{\text{ground}} = (0.77 \pm 0.01) \cdot 137\text{Cs}_{\text{shoulder}} \] with \( r^2 = 0.96. \)

In Table 5.7 are summarized the parameters of correlation between in-situ and laboratory measurements.

**Table 5.7:** correlation parameters between in-situ measurements on ground and on shoulder for two classes of vegetative coverage.

<table>
<thead>
<tr>
<th>Isotopes</th>
<th>( a \pm \sigma_a )</th>
<th>Vegetative coverage</th>
<th>Vegetative coverage</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td></td>
<td>0-50 %</td>
<td>50-100 %</td>
</tr>
<tr>
<td>Th [mg/Kg]</td>
<td>0.83 ( \pm 0.02 )</td>
<td>0.85 ( \pm 0.02 )</td>
<td></td>
</tr>
<tr>
<td>137 Cs [cps]</td>
<td>0.72 ( \pm 0.02 )</td>
<td>0.77 ( \pm 0.01 )</td>
<td></td>
</tr>
</tbody>
</table>

In the case of \( ^{232}\text{Th} \) it is observed a minor degree of influence due to the vegetative cover from 50-100%, but comparable within 1\( \sigma \) with the case of 0-50% of vegetative coverage. While, in the case of \( ^{137}\text{Cs} \) it is seen clearly within the 1\( \sigma \) the influence due to the presence of the vegetative cover. As it is expected, the presence of vegetative cover is more visible in the case of relatively lower energies.
Chapter 6

6. Conclusions

In this work, have been studied several performances of portable detector (ZaNaI_1.0L) composed by a solid scintillator NaI (Tl) of 1L of volume, designed and built specifically to measure the radioactivity content in soils and rocks for geological and environmental purposes. The aim of this thesis is the study of the performances of measurements carried out in-situ with the ZaNaI_1.0L by comparing them with measurements of the radioactivity content carried out in the laboratory on the soil samples representative of the same area investigated by the portable detector. Furthermore, since the ZaNaI_1.0L was designed to be used both in static and dynamic acquisition mode we have studied the performances of the measurements performed in different situations.

The area in which were chosen the sites in which are made the measurements of radioactivity is Ombrone basin, located in southern Tuscany and the Commune of Schio located in the Region of Veneto. During the campaign activity were investigated 80 and 89 sites respectively in Ombrone basin and Schio district with a total of 338 spectra. In addition, in each of the 80 sites investigated in Ombrone basin other than realizing a measurement of radioactivity with the ZaNaI_1.0L, were taken 5 different soil samples, with a total of 400 samples. The spectra acquired with ZaNaI_1.0L we analyzed using the FSA-NNLS method in order to determine the abundances of $^{40}$K, eU and eTh. Instead, the soil samples were measured in the laboratory with a high-pure germanium detector (MCA_Rad) and analyzed for $^{40}$K, eU and eTh abundances.

In particular, we have studied the correlation between:

- 80 spectra by placing the ZaNaI_1.0L on the ground in correspondence with 400 soil samples collected in order to characterize the site;
- 80 spectra by placing the ZaNaI_1.0L a tripod at 1m height in correspondence with 400 soil samples collected in order to characterize the site;
- 80 spectra by placing the ZaNaI_1.0L on a tripod at 1m height in correspondence with 80 spectra by placing the ZaNaI_1.0L on the ground;
• 89 spectra by placing the ZaNaI_1.0L on the shoulders of an operator in correspondence with 89 spectra by placing the ZaNaI_1.0L on the ground.

In general several sources of background radiation and environmental factors as: cosmic rays, atmospheric radon (\(^{222}\)Rn), rain and soil moisture, also the vegetative cover, can interfere the radiation emitted from potassium, uranium and thorium present in the soil and rocks during the measurements in-situ. In this study the influence due to changes on cosmic radiation are considered negligible since the sites are chosen at similar heights above sea-level. The effect of attenuation of gamma-rays due to changes in soil moisture after rainfalls is also considered negligible since all in-situ measurements are always performed at least 1 week after rainfalls. While the interference due to atmospheric radon for the signal coming from uranium it is discussed only qualitatively in our results. Finally, the attenuation effect due to vegetative cover is discussed and from the data it was possible to determine such interference only in the case of \(^{137}\)Cs, for the other isotopes the effect was found to be negligible within the uncertainty.

The comparison between in-situ measurements both on ground and on tripod (1m height) versus laboratory measurements shows a good correlation among them (Table 6.1).

**Table 6.1**: correlation between ZaNaI_1.0L measurements obtained by placing the detector on ground and on tripod respect to laboratory measurements.

<table>
<thead>
<tr>
<th>Isotope</th>
<th>([\text{ZaNaI}]_{\text{ground}} = (a \pm \sigma_a) \ [\text{HPGe}])</th>
<th>([\text{ZaNaI}]_{\text{tripod}} = (a \pm \sigma_a) \ [\text{HPGe}])</th>
</tr>
</thead>
<tbody>
<tr>
<td>(^{40})K (%)</td>
<td>1.16 ± 0.05</td>
<td>1.11 ± 0.05</td>
</tr>
<tr>
<td>eTh (µg/g)</td>
<td>0.85 ± 0.11</td>
<td>0.75 ± 0.10</td>
</tr>
<tr>
<td>eU (µg/g)</td>
<td>0.97 ± 0.12</td>
<td>0.92 ± 0.11</td>
</tr>
</tbody>
</table>

The final relative uncertainties for K, U and Th are found to be less than about 20% for ZaNaI_1.0L on ground versus HPGe measurements and about 35% ZaNaI_1.0L on tripod versus HPGe measurements.
The determination of the attenuation of air during in-situ measurements performed by placing the detector at 1m height (tripod) are important in order to report that the results at ground concentration of the radioactivity. For this purpose, we obtained this correction by studying the results in table 6.1 and by comparing directly in-situ measurement performed by placing the detector on-ground and on tripod. The results are found to be comparable within 1σ for U and eTh and out for more than 3σ for K. The results show a clear energy dependence of attenuation factor which decreases with increasing of gamma ray energies with exception of eU which seems to be influences by the presence of radon gas in air.

Table 6.2: attenuation correction due to air for in-situ measurements at tripod (1m height).

<table>
<thead>
<tr>
<th>Isotopes</th>
<th>Max. energy [keV]</th>
<th>air attenuation correction [%]</th>
</tr>
</thead>
<tbody>
<tr>
<td>$^{137}$Cs [cps]</td>
<td>662</td>
<td>19 ± 0.5</td>
</tr>
<tr>
<td>K [%]</td>
<td>1460</td>
<td>7 ± 0.3</td>
</tr>
<tr>
<td>U [mg/kg]</td>
<td>1674</td>
<td>13 ± 0.9</td>
</tr>
<tr>
<td>Th [mg/kg]</td>
<td>2615</td>
<td>6 ± 0.4</td>
</tr>
</tbody>
</table>

An important application of ZaNaI_1.0L is the utilization in dynamic mode of acquisition, where the detector configured in a back-pack in held on shoulder of an operator. In order to correct for the attenuation effect due to the presence of the operator we have compared the results obtained in-situ by placing the detector on ground and holding on shoulders (table 6.3). Again is can be observed the energy dependence of attenuation factor which decreases with increasing of gamma ray energies again with exception of eU which seems to be influences by the presence of radon gas in air.
Table 6.3: attenuation correction due to the presence of the operator during in-situ dynamics acquisition

<table>
<thead>
<tr>
<th>Isotopes</th>
<th>Max. energy [keV]</th>
<th>air + operator attenuation correction [%]</th>
</tr>
</thead>
<tbody>
<tr>
<td>$^{137}$Cs [cps]</td>
<td>662</td>
<td>23 ± 0.3</td>
</tr>
<tr>
<td>K [%]</td>
<td>1460</td>
<td>18 ± 0.2</td>
</tr>
<tr>
<td>U [mg/kg]</td>
<td>1674</td>
<td>16 ± 0.2</td>
</tr>
<tr>
<td>Th [mg/kg]</td>
<td>2615</td>
<td>17 ± 0.4</td>
</tr>
</tbody>
</table>
References

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